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# Insight into the effect of $SO_2$ on the $Hg^0$ removal performance over a 1V-8Ce/AC sorbent at low temperatures



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#### GRAPHICAL ABSTRACT



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#### ABSTRACT

The influences of  $SO_2$  on  $Hg^\circ$  removal over the 1V-8Ce/AC sorbent were systematically investigated at low temperatures. The experimental results showed that  $SO_2$  has a dual effect on  $Hg^\circ$  removal, that is,  $SO_2$  has both a promoting effect and an inhibiting effect on  $Hg^\circ$  removal. The  $SO_2$  transient response experiment indicated that  $SO_2$  could not only react with  $Hg^\circ$  to promote the removal of  $Hg^\circ$  but also react with the active components and poison the sorbent.  $O_2$  is indispensable for the removal of  $Hg^\circ$ , which can offset the adverse effects caused by  $SO_2$  and  $H_2O$ . HCl exhibited an obvious promoting effect on  $Hg^\circ$  removal in the presence of  $SO_2$ . The 1V-8Ce/AC sorbent exhibited good sulfur resistance and excellent stability ( $E_{Hg} = 90.04$ %) after a 24 h reaction performed under the 1000 ppm  $SO_2$  condition at 150 °C. In addition, the Hg-TPD and XPS methods were used to assist in studying the effect of  $SO_2$  on  $Hg^\circ$  removal over the 1V-8Ce/AC sorbent. Finally, the mechanism of  $Hg^\circ$  removal in an  $SO_2$  atmosphere was also explored, which showed that  $Hg^\circ$  was removed by two possible pathways over the 1V-8Ce/AC sorbent.

#### 1. Introduction

In recent years, with the rapid development of modern industry, environmental pollution has become increasingly prominent (Streets et al., 2011). However, with people's yearning for a better life,

environmental problems, especially air pollution, have been increasingly emphasized. Mercury, a natural trace element found in coal, is released after burning and mainly comes in three forms: elemental mercury (Hg<sup>0</sup>), oxidized mercury (Hg<sup>2+</sup>), and particulate mercury (Hg<sup>p</sup>) (Galbreath and Zygarlicke, 1996). Among them, Hg<sup>2+</sup> and Hg<sup>p</sup>

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can be easily controlled by air pollution control devices (APCDs), such as electrostatic precipitators (ESPs), fabric filters (FFs) and wet flue gas desulfurization equipment (WFGD) (Chen et al., 2018). However, only Hg<sup>0</sup> is difficult to be effectively removed due to its high volatility and poor water-solubility (Gao et al., 2013). Generally, selective catalytic reduction (SCR) is usually used in conjunction with a desulfurization unit, which results in a large investment and high operating costs. Furthermore, although Hg<sup>2+</sup> in flue gas can be effectively captured and enriched in slurry during the WFGD process, it can also cause secondary pollution (Xu et al., 2017; Gustin and Ladwig, 2010; Presto and Granite, 2006). Therefore, determining how to effectively control the release of Hg<sup>0</sup> into the air has become a hot and difficult research topic.

Currently, activated carbon (AC) is widely used as an sorbent due to its large specific surface area, developed pore structure, and great mechanical strength (Diamantopoulou et al., 2010; Wu et al., 2015; Xie et al., 2015). Previous studies have found that AC has a large specific surface area and has a certain adsorption effect on Hg°, after the modification of AC by V and Ce, it can significantly improve the removal performance of Hg° (Zhu et al., 2018). Although brominated activated carbon has shown excellent performance in the removal of Hg° (Saha, 2014; Zhou et al., 2015), the study of economic, environmental, and efficient sorbents (or catalysts) are still focused. By modifying AC, V-Ce/AC sorbent has both adsorption and oxidation properties on Hg°, which can further improve the removal performance of various pollutants without causing secondary pollution. What's more, the dry simultaneous removal technologies used to remove multi-pollutants from coal-fired flue gas by using carbon-based sorbents have achieved zero water consumption based on their high efficiency and low energy.

As a global pollutant, mercury has attracted increasing attention, and the emission requirements of pollutants emitted from coal-fired power stations have also become more strict (China, 2011; United States Environmental Protection Agency (U.S. EPA), 2013). Under such conditions, effectively controlling the removal of Hg<sup>0</sup> without causing secondary pollution has become a more environmentally friendly research method. As we all know, flue gas compositions, especially those containing acidic gases (such as HCl and SO<sub>2</sub>), have different effects on Hg<sup>0</sup> removal. It is generally believed that HCl improves Hg<sup>0</sup> removal due to the role of active chlorine species [Cl] (Tong et al., 2015; Li et al., 2012). However, the effect of SO<sub>2</sub> on Hg<sup>0</sup> removal has been controversial (Xu et al., 2017; Li et al., 2013), and the effect of SO<sub>2</sub> on the Hg<sup>0</sup> removal performance over a V-Ce/AC sorbent is rarely studied. Therefore, it is necessary to conduct a comprehensive and in-depth study on the effect of SO<sub>2</sub> on the Hg<sup>0</sup> removal performance over a V-Ce/AC sorbent

In this paper, a systematic study was performed to determine the influence of  $SO_2$  on the  $Hg^\circ$  removal performance over the 1V-8Ce/AC sorbent at low temperatures, which included studying the role of the concentration of  $SO_2$ , the effect of the temperature ( $100-220\,^\circ$ C), and the interaction of  $SO_2$  with other flue gas components (including  $O_2$ ,  $H_2O$ , and HCl). Moreover, the possible reaction pathways were also proposed using techniques such as XPS and Hg-TPD. The ultimate goal is to fully understand the mechanism of  $Hg^\circ$  removal under the  $SO_2$  atmosphere. This knowledge will play a guiding role in effectively controlling mercury emissions from coal-fired power plants.

## 2. Experimental

## 2.1. Chemicals and reagents

Activated carbon (AC) was supplied by Xinhua Chemical Ltd., China. All the reagents used in this work were analytical pure grade (AR), including: Cerium nitrate in hexahydrate (99.0 %, Sinopharm Chemical Reagent Co., Ltd), ammonium metavanadate (99.0 %, Kemiou Chemical Reagent Co., Ltd) and Oxalic acid (99.5 %, Sinopharm Chemical Reagent Co., Ltd). Ultrapure water was used for preparing the

required solutions.

#### 2.2. Catalyst preparation

AC was used as the catalyst support. The V-Ce/AC sorbent was synthesized by an ultrasound-assisted impregnation method. The details of a similar synthetic procedure are available in our previous research (Zhu et al., 2018).

#### 2.3. Catalyst characterization

The X-ray photoelectron spectroscopy (XPS) analysis was conducted with an Escalab 250Xi instrument (Thermo Fisher Scientific Corporation, USA) using monochromatic Al K-Alpha radiation as the excitation source. The spectra were corrected by the C 1s binding energy value of 284.6 eV.

The temperature-programmed desorption experiments of Hg (Hg-TPD) were carried out in a quartz reactor with 0.10 g of the samples (fresh or spent), which were heated from 60 °C to 660 °C with a heating rate of 5 °C/min and exposed to 80 mL/min pure  $N_2$ . The outlet Hg° concentration was continuously measured by an on-line RA-915 M mercury analyzer (Lumex Co., Ltd., Russia).

## 2.4. Catalyst activity test

As shown in Figure S1, the activity of the catalyst was tested using a bench-scale fixed-bed reactor that consists of several parts, including the flue gas simulation system, temperature control system, reaction system, online analysis system, and exhaust treatment system. During each test, 0.10 g of the samples were packed into the quartz reactor. The total flow rate was 240 mL/min. All the simulated flue gas components, including  $N_2$ ,  $O_2$ ,  $SO_2$ ,  $H_2O$ , and HCl, were precisely controlled by a mass flow controller (MFC). Among these flue gas components, the  $N_2$  flow was used for two purposes: one purpose was used to carry water vapor into the flue gas system, and another purpose was to carry the  $Hg^\circ$  vapor. Before and after each test, the concentrations of the feed gases were precisely detected.

The concentrations of  $Hg^{\circ}$  and  $SO_2$  were detected by an RA-915 M mercury analyzer and a KM9506 flue gas analyzer (KANE International Limited, U.K), respectively.

The  $Hg^0$  removal efficiency ( $E_{Hg}$ ) is defined as follows:

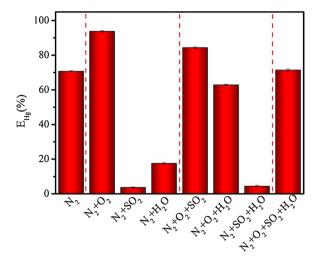
$$E_{Hg}(\%) = \frac{Hg_{in}^{0} - Hg_{out}^{0}}{Hg_{in}^{0}} \times 100\%$$
 (1)

where  $Hg_{out}^0$  and  $Hg_{out}^0$  are the inlet and outlet  $Hg^0$  concentrations of the fixed-bed reactor, respectively ( $\mu g/m^3$ ).

#### 3. Results and discussion

## 3.1. Effect of different flue gas components on Hg° removal

The Hg° removal performance of the 1V-8Ce/AC sorbent was investigated under the different flue gas component conditions. It can be clearly seen from Fig. 1 that the  $E_{\rm Hg}$  varies significantly under the different atmospheres. The  $E_{\rm Hg}$  in pure  $N_2$  was only 70.59 %. After 6 vol %  $O_2$  was added into the gas stream, the  $E_{\rm Hg}$  obviously increased to 93.76 %. However, when 1000 ppm SO $_2$  or 8 vol%  $H_2O$  was added into pure  $N_2$ , which showed an obvious downward trend and only 3.64 % or 17.55 % was achieved, respectively. This may be due to the competitive adsorption between Hg° and SO $_2$  (or  $H_2O$ ) (Wen et al., 2011; Xu et al., 2015). After 1000 ppm SO $_2$  was added into the  $N_2+O_2$  atmosphere, the  $E_{\rm Hg}$  declined to 84.35 %. Similarly, when 8 vol%  $H_2O$  was added into the  $N_2+O_2$  atmosphere, the  $E_{\rm Hg}$  declined to 62.80 %. However, when 1000 ppm SO $_2$  and 8 vol%  $H_2O$  were added together into pure  $N_2$ , the  $E_{\rm Hg}$  seriously declined to only 4.37 %. After 6 vol%  $O_2$  was added into



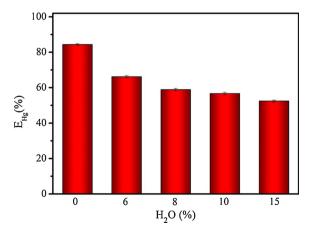
**Fig. 1.** Effect of different flue gas components on Hg $^{\circ}$  removal over the 1V-8Ce/AC sorbent at 150  $^{\circ}$ C (reaction conditions: O<sub>2</sub> = 6 vol $^{\circ}$ M, Hg $^{\circ}$  = 240  $\mu$ g/m $^{3}$ , SO<sub>2</sub> = 1000 ppm, H<sub>2</sub>O = 8 vol $^{\circ}$ M, N<sub>2</sub> as balance, GHSV = 80,000 h $^{-1}$ ).

the  $N_2 + SO_2 + H_2O$  atmosphere, the  $E_{Hg}$  increased back to 71.39 %. This increase indicated that  $SO_2$  and  $H_2O$  could inhibit the removal of  $Hg^\circ$ , while  $O_2$  had an obvious promoting effect on  $Hg^\circ$  removal, and the addition of  $O_2$  could offset the inhibition of  $SO_2$  and  $H_2O$  (Zhu et al., 2018; Zhang et al., 2014).

Based on the abovementioned studies about the effects of the various gases on Hg $^\circ$  removal, an O $_2$  transient response experiment was conducted over the 1V-8Ce/AC sorbent in the presence of SO $_2$ . As shown in Figure S2, the E $_{\rm Hg}$  of the 1V-8Ce/AC sorbent in the simulated flue gas (Hg $^\circ$ +H $_2$ O + SO $_2$ ) without added O $_2$  was maintained at approximately 5% for 60 min. However, when 6 vol% O $_2$  was added to the simulated flue gas, the E $_{\rm Hg}$  increased rapidly and remained at approximately 66 %. When 6 vol% O $_2$  was cut off again at the 180th min, the E $_{\rm Hg}$  decreased rapidly from 64.13 % to 6.19 % within 20 min and remained at very low efficiency. The results indicated that O $_2$  can offset the adverse effects caused by SO $_2$  and H $_2$ O in the removal of Hg $^\circ$ .

## 3.2. Effect of H<sub>2</sub>O concentrations on Hg° removal

Water vapor, an important component in flue gas, was also investigated under the 1000 ppm  $SO_2$  condition. As shown in Fig. 2, adding 6 vol%  $H_2O$  into the simulated flue gas significantly impacted the activities of the 1V-8Ce/AC sorbent and caused the  $E_{Hg}$  to significantly declined from 84.35% to 66.18%, and the  $E_{Hg}$  continued to



**Fig. 2.** Effects of different  $\rm H_2O$  concentrations on  $\rm Hg^\circ$  removal at 150 °C (reaction conditions:  $\rm O_2=6$  vol%,  $\rm Hg^\circ=240~\mu g/m^3, H_2O=0$ -15 vol%,  $\rm SO_2=1000~ppm, N_2$  as balance, GHSV = 80,000 h $^{-1}$ ).

decreased from 58.89~% to 52.43~% with increasing the  $H_2O$  content (from 8~vol% to 15~vol%). This indicated that the addition of  $H_2O$  has an obvious inhibitory effect on the removal of  $Hg^\circ$ , and the inhibitory effect is more obvious with increasing  $H_2O$  contents.

To comprehensively understand the relationship between  $\rm H_2O$  and  $\rm Hg^\circ$  under the  $\rm SO_2$  atmosphere condition, the transient response of  $\rm H_2O$  was evaluated. As seen from Figure S3, the  $\rm E_{Hg}$  significantly decreased from 89.98 % to 70.40 % after 8 vol%  $\rm H_2O$  was added at the 60th min, but the  $\rm E_{Hg}$  could be recovered to 89.59 % when 8 vol%  $\rm H_2O$  was removed from the simulated flue gas system at the 120th min. This indicated that there is an obvious competitive adsorption relationship between  $\rm H_2O$  and  $\rm Hg^\circ$  on the surface of the 1V-8Ce/AC sorbent (Zhang et al., 2014; Zhou et al., 2016), and this temporary deactivation can be stopped and the original level restored when the  $\rm H_2O$  supply was cut off.

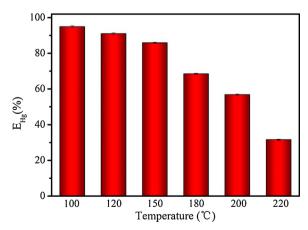
#### 3.3. Effect of temperature on Hg° removal in the presence of SO<sub>2</sub>

To investigate the effect of different temperatures on Hg $^{\circ}$  removal in the presence of SO<sub>2</sub>, 1000 ppm SO<sub>2</sub> was added into the simulated flue gas, and the experiments were conducted at a temperature range of 100-220 °C.

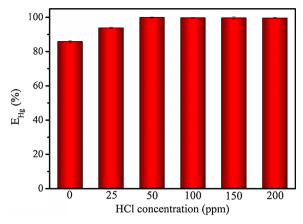
As shown in Fig. 3, when the reaction temperature was adjusted to 100 °C, the highest removal efficiency was obtained ( $E_{Hg} = 94.93$  %). By continuing to increase the reaction temperature to 120 °C, 150 °C, 180 °C, and 200 °C, the obtained  $E_{Hg}$  was 91.01 %, 85.87 %, 68.5 %, and 56.84 %, respectively. Note that the  $E_{\mbox{\scriptsize Hg}}$  dropped significantly, and an E<sub>Hg</sub> of only 31.58 % was obtained after further increasing the temperature to 220 °C. This indicated that the reaction temperature has an obvious influence on the Hg° removal reaction, and higher temperatures were more unfavorable for Hg° removal under the SO<sub>2</sub> condition. According to the abovementioned experimental results, two possible reasons might explain this inhibition trend. The first reason is that the temperature increase is not conducive to the adsorption of Hg° (limiting the first step of the Hg° removal reaction) (Zhu et al., 2019). The second reason is that the reaction temperature increase accelerated the toxicity of the catalyst, where SO2 could react with the active components to form sulfate and the caused partial deactivation of the 1V-8Ce/AC sorbent (Xu et al., 2015; Ma et al., 2008; He et al., 2016a).

## 3.4. Effect of HCl on $Hg^{\circ}$ removal in the presence of $SO_2$

As shown in Fig. 4, the effect of HCl on Hg $^\circ$  removal over the 1V-8Ce/AC sorbent at 150  $^\circ$ C was experimentally studied. When 25 ppm HCl was added to the simulated flue gas, the  $E_{\rm Hg}$  increased from 85.87% to 93.82%. After that, when the concentration of HCl was further



**Fig. 3.** Effect of temperature on Hg $^\circ$  removal in the presence of SO $_2$  (reaction conditions: O $_2=6$  vol%, Hg $^\circ=240\,\mu\text{g/m}^3$ , SO $_2=1000$  ppm, N $_2$  as balance, T = 100-220  $^\circ$ C, GHSV = 80,000 h $^{-1}$ ).



**Fig. 4.** Effect of different HCl concentrations over the 1V-8Ce/AC sorbent (reaction conditions:  $O_2=6$  vol%,  $Hg^\circ=240~\mu g/m^3$ , HCl=0-200~ppm,  $SO_2=1000~ppm$ ,  $N_2$  as balance,  $T=150~^\circ C$ ,  $GHSV=80,000~h^{-1}$ ).

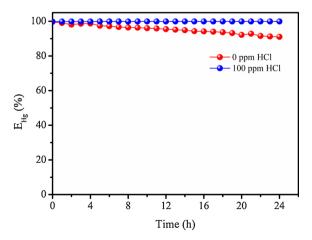
increased to 200 ppm, the  $E_{Hg}$  almost remained at 99.93 %. This indicated that HCl has an obvious promoting effect on Hg $^{\circ}$  removal (Niu et al., 2015).

Subsequently, we continued to study the lifetime of the 1V-8Ce/AC sorbent under the 100 ppm HCl condition. The results are shown in Fig. 5. For the reaction without added HCl, the  $E_{\rm Hg}$  of 1V-8Ce/AC gradually decreased as the experiment time was prolonged, and the  $E_{\rm Hg}$  decreased to 91 % after 24 h. However, for the reaction where 100 ppm HCl was added to the simulated flue gas, the  $E_{\rm Hg}$  of 1V-8Ce/AC was consistently maintained at approximately 99 % after 24 h. This indicated that the addition of HCl contributed to the stability of Hg $^{\circ}$  removal. In the presence of HCl and  $O_2$ , HCl can be oxidized to form active chlorine species that could react with Hg $^{\circ}$  to form HgCl $_2$  (Wen et al., 2011; Xu et al., 2015; Chen et al., 2016). Thus, the pathway for the reaction of HCl and Hg $^{\circ}$  over the 1V-8Ce/AC sorbent might be described as Eq. (2).

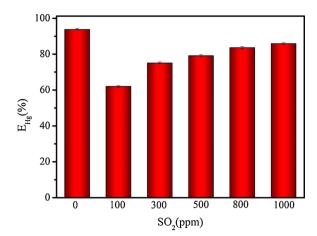
$$4HCl(ads) + O_2(ads) + 2Hg^0(ads) \rightarrow 2HgCl_2(ads) + 2H_2O(ads)$$
 (2)

## 3.5. Effect of SO<sub>2</sub> concentrations on Hg° removal

The effect of  $SO_2$  concentrations on  $Hg^\circ$  removal is shown in Fig. 6. It can be seen that the  $E_{Hg}$  of 1V-8Ce/AC was 93.76 % without added  $SO_2$ . When 100 ppm  $SO_2$  was added into the simulated flue gas stream, there was an obvious inhibitory trend, and the  $E_{Hg}$  decreased to 55.25 %. However, an interesting experimental phenomenon was observed, as



**Fig. 5.** The stability of 1V-8Ce/AC sorbent under the HCl atmosphere at 150 °C (reaction conditions:  $O_2=6\%$ ,  $Hg^\circ=240~\mu g/m^3$ , HCl=100~ppm,  $N_2$  as balance,  $GHSV=80,000~h^{-1}$ ).



**Fig. 6.** Effect of different SO<sub>2</sub> concentrations on Hg $^\circ$  removal at 150  $^\circ$ C (reaction conditions: O<sub>2</sub> = 6 vol%, Hg $^\circ$  = 240 µg/m $^3$ , SO<sub>2</sub> = 0-1000 ppm, N<sub>2</sub> as balance, GHSV = 80,000 h $^{-1}$ ).

the  $SO_2$  concentration in the flue gas flow increased to 300 ppm, 500 ppm, and 800 ppm, the  $E_{Hg}$  gradually increased to 75.04 %, 79.09 %, and 83.61 % again, respectively. By further increasing the  $SO_2$  concentration to 1000 ppm, the  $E_{Hg}$  was increased to 85.87 %. This indicated that  $SO_2$  has a certain inhibitory effect on  $Hg^\circ$  removal, but the inhibition trend slowed down with increasing  $SO_2$  concentrations. The extent to which the  $SO_2$  concentration affected  $Hg^\circ$  removal differed depending on the  $SO_2$  concentration, which might be attributed to  $SO_2$  adsorbed on the surface of the sorbent, and a small part of  $SO_2$  is oxidized to  $SO_3$ . With increasing  $SO_2$  concentrations, the amount of  $SO_2$  that oxidized to accumulated  $SO_3$  gradually increases, and  $SO_3$  can react with  $Hg^\circ$  to generate  $HgSO_4$  (Zhang et al., 2017; Li et al., 2011; Fan et al., 2012), thus reducing the extent to which  $SO_2$  inhibited the  $SO_3$  removal process to a certain degree.

To further investigate the effect of  $SO_2$  on the  $Hg^0$  removal performance, the  $SO_2$  transient response over the 1V-8Ce/AC sorbent and blank AC sample was evaluated experimentally. The results are shown in Fig. 7.

The  $SO_2$  transient response experiment was first carried out over the blank AC and then performed with the 1V-8Ce/AC sorbent for comparison. It can be clearly seen from Fig. 7 that the  $E_{\rm Hg}$  of the AC gradually decreased and was maintained at 4.90 % until 60 min. After 1000 ppm  $SO_2$  was added into the gas stream at the 60th min, a promotion trend was observed, and the  $E_{\rm Hg}$  increased from 4.90 % to 24.53 %. However, when the supply of  $SO_2$  was cut off at the 120th min, the  $E_{\rm Hg}$  decreased, but the  $E_{\rm Hg}$  (approximately 11.98 %) was still higher

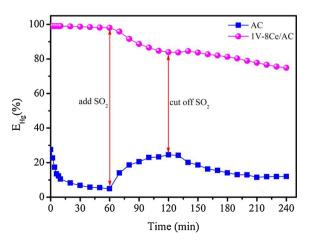


Fig. 7. Transient response of SO<sub>2</sub> (reaction conditions:  $O_2=6$  vol%,  $Hg^\circ=240$   $\mu g/m^3$ ,  $SO_2=1000$  ppm,  $N_2$  as balance, T=150 °C, GHSV = 80,000  $h^{-1}$ ).

than that observed before the addition of SO<sub>2</sub>. This indicated that the addition of SO<sub>2</sub> is helpful for the removal of Hg°, and the SO<sub>2</sub> remaining on the surface of AC can continue to react with Hg°.

Similarly, the  $SO_2$  transient response experiment was also conducted over the 1V-8Ce/AC sorbent. For the first 60 min, the  $E_{Hg}$  of the 1V-8Ce/AC sorbent was maintained at 98 % without added  $SO_2$ . When 1000 ppm  $SO_2$  was added to the simulated flue gas at the 60th min, the  $E_{Hg}$  gradually decreased from 98.06% to 83.74%. However, the  $E_{Hg}$  was not restored but continued to declined to 74.84 % after 1000 ppm  $SO_2$  was removed at the 120th min. Therefore, it can be inferred that the inhibitory role of  $SO_2$  on mercury removal is due to the partial deactivation caused by the toxicity of  $SO_2$  on the 1V-8Ce/AC sorbent. More importantly, it also can be seen that the toxicity of  $SO_2$  on the 1V-8Ce/AC sorbent is irreversible.

By comparing the two abovementioned experimental groups, we can infer that the added  $SO_2$  can react with  $Hg^0$ , which thus promoted  $Hg^0$  removal to a certain extent (Li et al., 2013). The AC modified by active components ( $V_2O_5$  and  $CeO_2$ ) further improved the removal performance of  $Hg^0$ . However, the added  $SO_2$  can also react with the active components of the catalyst, so the catalyst gradually lost its original activity. This may be due to the formation of metal sulfate during the reaction (Ma et al., 2008; Smirnov et al., 2014; He et al., 2016b). The process can be described as the following reaction in Eqs. (3) and (4).

$$2V_2O_5 + 4SO_2 + O_2 \rightarrow 4VOSO_4 \tag{3}$$

$$2\operatorname{CeO}_2 + 3\operatorname{SO}_2 + \operatorname{O}_2 \rightarrow \operatorname{Ce}_2(\operatorname{SO}_4)_3 \tag{4}$$

To determine the chemical state of the different samples (fresh and spent 1V-8Ce/AC), the XPS analysis was conducted, and the S 2p and Hg 4f spectra are shown in Fig. 8.

The S 2p spectra are shown in Fig. 8(A). For the fresh 1V-8Ce/AC sorbent, no peak was attributed to S 2p. In contrast, for the spent 1V-8Ce/AC sorbent spectrum, there were two obvious peaks at approximately 164.6 eV and 169.15 eV, which can be attributed to SO<sub>2</sub><sup>2</sup> species, indicating that SO<sub>2</sub> can be oxidized to SO<sub>2</sub><sup>2</sup> (Ma et al., 2008; Yang et al., 2017). The spectra of Hg 4f are presented in Fig. 8(B); the two main peaks at approximately 101.75 eV and 102.25 eV corresponded to Si and Hg<sup>2+</sup>, indicating that Hg<sup>2+</sup> formed on the spent 1V-8Ce/AC sorbent under the SO<sub>2</sub> condition (Zhu et al., 2018; Zhang et al., 2017). However, the Hg° peak was not observed from the Hg 4f spectra, which might be due to the Hg° content on the spent 1V-8Ce/AC sorbent surface being lower than the detection limit of the XPS analysis (Wang et al., 2016).

As we all know, different mercury species have different boiling points. To obtain a better understanding of the mechanism of how  $SO_2$  affects the  $Hg^0$  removal process, the Hg-TPD profile is an important consideration, as Hg-TPD can be used to effectively identify the type of mercury species in samples. By comparing the decomposition temperature of the mercury species in the samples with that of the pure

mercury compounds, the mercury species in the samples can be identified (Yang et al., 2017).

The Hg-TPD-O2 experiment was carried out on the spent 1V-8Ce/AC samples (treated with 6 vol%  $O_2 + 240 \mu g/m^3 Hg^\circ$ ). As shown in Fig. 9(A), with increasing temperatures, different types of mercury species were desorbed from the spent 1V-8Ce/AC sorbent, and two peaks appeared at approximately 230 °C and 380 °C. The peak at 230 °C was attributed to chemical adsorbed Hg° and the peak at 380 °C can be ascribed to oxidized mercury (Zhou et al., 2017; Rumayor et al., 2015; Sun et al., 2017). It can be inferred that there are two types of mercury species (Hg° and HgO) in the spent 1V-8Ce/AC sorbent, and it is also confirmed that the mercury removal process is the result of adsorption and oxidation. As for the Hg-TPD-O<sub>2</sub> + SO<sub>2</sub> experiment performed on the spent 1V-8Ce/AC samples (treated with 6 vol% O<sub>2</sub> + 1000 ppm SO<sub>2</sub> + 240 μg/m<sup>3</sup> Hg°) Fig.9(B) shows that three peaks can be observed at 250 °C, 370 °C, and 520 °C, respectively. The peak at 250 °C was attributed to chemical adsorbed Hg° and the peak at 370 °C can be classified as HgO. However, note that a peak at 520 °C is observed in Fig. 9(B) but not in Fig. 9(A). This new peak may be due to the formation of HgSO<sub>4</sub> on the surface of the catalyst after the addition of SO<sub>2</sub> during the reaction process (Tong et al., 2015; Zhou et al., 2017; Lopez-Anton et al., 2010). Based on the abovementioned experiments, we can infer the possible reactions as follows:

$$Hg^0(g) \rightarrow Hg^0(ads)$$
 (5)

$$O_2(ads) + Hg^0(ads) \rightarrow HgO(ads)$$
 (6)

$$O_2(ads) + SO_2(ads) + Hg^0(ads) \rightarrow HgSO_4(ads)$$
 (7)

## 3.6. Mechanism of the Hg<sup>0</sup> removal process in the presence of SO<sub>2</sub>

Combined with the abovementioned experimental and characterization results (Transient response of  $SO_2$ , Hg-TPD and XPS), the  $Hg^0$  removal reaction with  $SO_2$  over the 1V-8Ce/AC sorbent mainly consists of adsorption and oxidation, which can be divided into two pathways. Path 1. Inhibiting effect pathway

The first pathway is the inhibition effect on  $Hg^0$  removal (including the competitive adsorption and poisoning of the active sites). The inhibiting effect of water vapor may be due to the competitive adsorption of  $Hg^0$  and  $H_2O$  on the sorbent surface, which is temporarily deactivated and can be restored when water vapor is removed from the simulated flue gas, the reaction is shown in Eq. (8).

$$H_2O(g) \rightarrow H_2O(ads)$$
 (8)

However, the inhibiting effect of  $SO_2$  may be caused by the reaction between the active components (V and Ce) and  $SO_2$  (Ma et al., 2008; Smirnov et al., 2014; He et al., 2016b), so the sorbent gradually deactivates, and that's irreversible. The reaction is shown in Eqs. (3) and (4).

$$2V_2O_5 + 4SO_2 + O_2 \rightarrow 4VOSO_4$$
 (3)

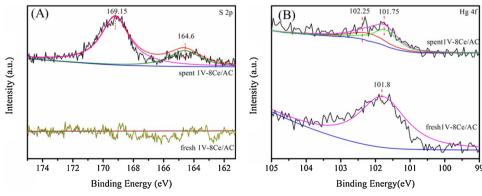


Fig. 8. S 2p (A) and Hg 4f (B) XPS spectra obtained for the fresh and spent 1V-8Ce/AC sorbent.

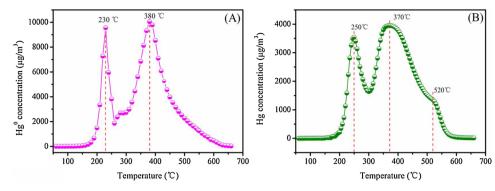
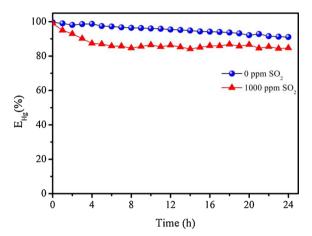


Fig. 9. Hg-TPD profiles: (A) Hg-TPD-O<sub>2</sub> and (B) Hg-TPD-O<sub>2</sub>+SO<sub>2</sub> (reaction conditions: N<sub>2</sub> = 80 mL, 5 °C/min, T = 60-660 °C).



**Fig. 10.** Stability of the 1V-8Ce/AC sorbent in the presence of  $SO_2$  (reaction conditions:  $O_2=6$  vol%,  $Hg^\circ=240~\mu g/m^3$ ,  $SO_2=1000$  ppm,  $N_2$  as balance,  $T=150~^\circ C$ ,  $GHSV=80,000~h^{-1}$ ).

$$2\operatorname{CeO}_2 + 3\operatorname{SO}_2 + \operatorname{O}_2 \rightarrow \operatorname{Ce}_2(\operatorname{SO}_4)_3 \tag{4}$$

Path 2. Promoting effect pathway

The second pathway is the promotion effect on Hg<sup>0</sup> removal (including adsorption and oxidation), as shown in Eqs. (2),(5),(6), and (7).

$$4HCl(ads) + O_2(ads) + 2Hg^0(ads) \rightarrow 2HgCl_2(ads) + 2H_2O(ads)$$
 (2)

$$Hg^0(g) \rightarrow Hg^0(ads)$$
 (5)

$$O_2(ads) + Hg^0(ads) \rightarrow HgO(ads)$$
 (6)

$$O_2(ads) + SO_2(ads) + Hg^0(ads) \rightarrow HgSO_4(ads)$$
 (7)

In the actual reaction process, the abovementioned two pathways will occur simultaneously. Therefore, the experimental phenomenon we finally observed is a balance of the results, in which the two pathways occurred simultaneously. What's more, the promoting effect is greater than the inhibiting effect in the presence of  $SO_2$  at the low reaction temperature (below 200 °C). Therefore, the 1V-8Ce/AC sorbent exhibits good sulfur resistance and excellent mercury removal performance.

## 3.7. The stability of the 1V-8Ce/AC sorbent under the SO<sub>2</sub> atmosphere

Based on the abovementioned speculation, the experimental results we finally observed are the results of the two reaction pathways equilibrating. Thus, the stability of the 1V-8Ce/AC sorbent during the Hg $^\circ$  removal process was also studied. As shown in Fig. 10, for the reaction without added SO $_2$ , the E $_{\rm Hg}$  gradually decreased, and an E $_{\rm Hg}$  of 91.06 % was obtained after 24 h. For the reaction with 1000 ppm SO $_2$  added, the E $_{\rm Hg}$  dropped to approximately 85 % during the first 4 h and gradually stabilized as the reaction time was prolonged. However, the E $_{\rm Hg}$  was

still maintained at approximately 84.68% after  $24\ h$ . This indicated that the 1V-8Ce/AC sorbent has a good sulfur resistance and excellent stability.

#### 4. Conclusions

In summary, the 1V-8Ce/AC sorbents were prepared by the ultrasound-assisted impregnation method and systematically investigated the influences of  $SO_2$  on  $Hg^\circ$  removal at low temperatures. The experimental results indicated that  $SO_2$  has both a promoting effect and an inhibiting effect on  $Hg^\circ$  removal. In addition, the effect of other flue gas components  $(O_2, H_2O,$  and HCl) on  $Hg^\circ$  removal was also investigated.  $O_2$  is indispensable for the removal of  $Hg^\circ$ , which can offset the adverse effects caused by  $SO_2$  and  $H_2O$ . HCl has an obvious promoting effect on  $Hg^\circ$  removal even in the presence of  $SO_2$ . The 1V-8Ce/AC sorbent exhibited good sulfur resistance and excellent stability  $(E_{Hg} = 90.04 \%)$  in the presence of the 1000 ppm  $SO_2$  at 150 °C. The  $Hg^\circ$  removal mechanism in the presence of  $SO_2$  was also explored by Hg-TPD and XPS methods, which showed that  $Hg^\circ$  was removed by two possible pathways over the 1V-8Ce/AC sorbent.

## CRediT authorship contribution statement

Youcai Zhu: Methodology, Formal analysis, Investigation, Writing - original draft, Data curation. Caiting Li: Supervision, Writing - review & editing. Yue Lyu: Writing - review & editing. Shanhong Li: Supervision, Writing - review & editing. Yindi Zhang: Writing - review & editing. Xueyu Du: Writing - review & editing. Yunbo Zhai: Supervision.

# **Declaration of Competing Interest**

The authors declare that there are no conflicts of interest.

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## Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:https://doi.org/10.1016/j.jhazmat.2020.123502.

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