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# Recent advances of photocatalytic degradation for BTEX: Materials, operation, and mechanism

Caixia Liang <sup>a,b</sup>, Caiting Li <sup>a,b,\*</sup>, Youcai Zhu <sup>a,b</sup>, Xueyu Du <sup>a,b</sup>, Chaoliang Yao <sup>c</sup>, Ying Ma <sup>c</sup>, Jungang Zhao <sup>a,b</sup>

<sup>a</sup> College of Environmental Science and Engineering, Hunan University, Changsha 410082, PR China

<sup>b</sup> Key Laboratory of Environmental Biology and Pollution Control (Hunan University), Ministry of Education, Changsha 410082, PR China

<sup>c</sup> Yonker Environmental Protection Co., Ltd, Changsha 410330, PR China

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# ABSTRACT

BTEX with benzene ring and different groups (methyl and ethyl) have aroused enormous concern due to their wide sources and large contribution to photochemical reactions and ozone pollution. This paper systematically reviewed the photodegradation of BTEX including materials, operating parameters, and reaction mechanisms. The photocatalytic materials including monadic or binary metal oxides, hydroxides, and metal-free materials were broadened the optical absorption range and enhanced the absorption capacity by modifying, thus greatly improving the photocatalytic performance. The most effective strategy is forming the heterojunction with builtin fields, which enhances the light absorption capacity and leads lower recombination rate of photocarrier to generate more active oxygen species. The parameters of operating mode, initial concentration, flow rate, temperature, relative humidity and light play important roles in photocatalytic degradation of BTEX. Furthermore, this work highlights that the process of photo-oxidizing BTEX overwhelmingly follows the Langmuir-Hinshelwood model. The mechanisms including adsorption and surface reaction are described using the density functional theory calculations. The review reveals that BTEX are oxidized to the aldehyde and organic acids and opened the ring by active species, eventually mineralized to CO<sub>2</sub> and H<sub>2</sub>O. Moreover, the different degradation pathways of BTEX are studied with the differences of position and number of methyl and ethyl groups. The technique combination is mostly used in the removal of BTEX due to complex working conditions in practical application. The challenge of catalyst inactivation can be regenerated by heating and UV cleaning.

## 1. Introduction

Recent decades have seen a growing tendency in the emissions of volatile organic compounds (VOCs), which has caused a series of air pollutions for their mobility and instability [1]. Anthropogenic VOCs, especially various industrial emissions [2], are precursors of ozone and photochemical smog, which are harmful to the environment and human health due to their toxicity, such as carcinogenicity, mutagenicity, and teratogenicity [3–6]. The vast majority of emitted VOCs consist of al-kanes, alkenes, alkynes, aromatics, alcohols, aldehydes, ketones, esters, and halocarbons. Aromatics are recognized as high pollutants due to their large contribution to photochemical reactions and ozone pollution. Aromatic compounds (BTEX, benzene, toluene, ethylbenzene, and xylene), have a wide range of sources, such as automobile exhaust and organic solvents in building decoration materials, which have attracted

great public attention [7]. BETX with different structures with benzene ring contribute to the majority of the total industrial emissions, which facilitate research and comparison. It is greatly worth noting the promising technologies for removing the aromatic compounds to protect the environment and human beings.

Over the past few decades, there are numerous abatement technologies to degrade BTEX, including absorption [8], adsorption [9], membrane separation [10], photocatalytic oxidation [11], catalytic oxidation [12], plasma destruction [13], and biological oxidation [14], etc. Among them, photocatalytic technology has gained increasing attention due to the highly efficient, low-energy, and harmlessness, which has broad application prospects in many fields [15,16]. At the background of conventional energy crisis, photocatalytic technology become the hot spot. Photocatalytic materials convert light energy into chemical energy under ultraviolet or visible light, and promote the

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<sup>\*</sup> Corresponding author at: College of Environmental Science and Engineering, Hunan University, Changsha 410082, PR China. *E-mail address*: ctli@hnu.edu.cn (C. Li).

synthesis or decomposition of organic materials [17]. When the light energy is equal or exceeds to the bandgap energy of the semiconductor material, the electrons are excited from the valence band (VB) to the conduction band (CB) to form photo-generated carriers (electron-hole pairs), in which valence band holes are strong oxidants, while conduction band electrons are strong reductants [18,19]. Photogenerated electron-hole pairs can react with O<sub>2</sub> or H<sub>2</sub>O adsorbed on the catalyst surface to produce superoxide radicals or hydroxyl radicals. These strongly active radicals can oxidize BTEX to mineralize to CO<sub>2</sub> and H<sub>2</sub>O eventually [20].

Recently, some reviews on photocatalytic oxidation of VOCs have been published. The work of Shayegan's group has focused on the strategies for improving TiO2 photocatalyst activities in VOCs degradation [21]. Vikrant et al. [22] spotlighted on the photocatalytic systems for benzene removal including the continuous mode and the batch mode. Zou et al. [23] summarized recent investigations on the integration of adsorption and photocatalytic degradation of VOCs using carbonbased nanocomposites. Besides, the key challenges for photocatalytic air purification to improving the applicability, including visible light utilization and catalyst deactivation have also been reported by Weon's group [24]. Nevertheless, to the best of our knowledge, little attention has been paid on the systematical review of the photocatalytic degradation of BTEX focusing on the mechanism of photocatalysis and discussing the practical application of BTEX photodegradation technology. In this review, we provide a detailed summary on the recent advances for BTEX removal using photocatalytic technologies. BTEX, as typical VOCs with special structure, can be systematically studied to understand the photocatalytic mechanism deeply and improve the practical application. Firstly, it provides an overview of photocatalytic materials including titanium dioxide, carbon nitride, zinc oxide and Bi-based photocatalyst, which have been mostly applied in BTEX removal. In addition, the operating parameters had a profound impact on the removal efficiency of BTEX. Furthermore, the kinetic model and mechanisms of photocatalytic decomposition of BTEX are described in detail. Finally, this article discusses the practical applications to combine other

technologies with photocatalytic technology about the VOCs photodegradation and suggests some methods to rebirth the deactivated catalysts, which greatly improves the practical application of photocatalytic technology.

# 2. Materials for BTEX photocatalysis

# 2.1. TiO<sub>2</sub>-based photocatalysts

Considering the economic and environmental benefits, TiO<sub>2</sub> is vigorously applied to the field of photocatalysis attributing to its excellent catalytic performance, stable chemical performance, safety, non-toxicity, and long service life [25,26]. In general, there are three crystal structures of  $TiO_2$  in nature: rutile (Eg = 3.02 eV), anatase (Eg = 3.2 eV), and brookite (Eg = 3.14 eV). Among them, anatase TiO<sub>2</sub> has the best catalytic activity [27]. However, the application of TiO<sub>2</sub> is astricted by the large energy band gap, leading to only 4 % of the solar light utilization rate [24,28]. Besides, the development of TiO<sub>2</sub> is also hindered by the high rate of charge carrier recombination and insensitivity to visible light in photocatalysis (Fig. 1a). Accordingly, multifarious strategies have been developed vigorously, which include doping with metals and non-metals elements, and coupling with some semiconductors. The modified TiO2 catalysts obtained by different methods exhibited better performance for photodegradation of BTEX, which listed in the Table 1.

## 2.1.1. Doping with metals

Noble metals play a major role in various organic conversions. Noble metal modified  $TiO_2$  can greatly accelerate the separation of photogenerated electron-hole pairs significantly improving the photocatalytic activity [56,57]. Noble metals can promote the absorption of  $TiO_2$  in visible light [58]. However, the use of noble metals for catalysis is greatly limited by their price. Non-noble metals (Mn, Fe, Cu, Ce, etc.) are often used as substitutes for modified  $TiO_2$  due to their low price and wide source [59–61]. Non-noble metals can also narrow the band gap of



**Fig. 1.** (a) UV–vis diffuse reflectance spectra of  $TiO_2$  thin film on glass [16]. (b) UV–vis diffuse reflection spectra of  $TiO_2$  modified with non-noble metals [51]. (c) The removal efficiency (RE) of gaseous benzene on  $TiO_2$  modified with various base metals [51]. (d) The RE of gaseous toluene on  $TiO_2$  modified with different dopants metals [52–55].

#### Table 1

The photodegradation of BTEX using TiO<sub>2</sub>-based photocatalysts.

Photocatalyst	Method	Light	BETX	Concentration	RE <sup>a</sup> (%)	Reference
P25	wet impregnation	UV	benzene	2 ppm	20	[29]
			toluene	2 ppm	15	
			ethylbenzene	2 ppm	50	
			xylene	2 ppm	82	
Pt-P25	wet impregnation	UV	benzene	2 ppm	7	
			toluene	2 ppm	31	
			ethylbenzene	2 ppm	81	
			xylene	2 ppm	72	
Fe-P25	wet impregnation	UV	benzene	2 ppm	12	
	1 0		toluene	2 ppm	19	
			ethylbenzene	2 ppm	74	
			xylene	2 ppm	81	
Ce-P25	wet impregnation	UV	benzene	2 ppm	80	
60-125	iver impregnation		toluene	2 ppm	26	
			ethylbenzene	2 ppm	85	
			xylene	2 ppm	78	
V-TiO <sub>2</sub>	solvothermal	visible	benzene	1000–1200 ppm	30	[30]
v-110 <sub>2</sub>	sorvotricimar	VISIDIC	toluene	1000–1200 ppm	68	[30]
			ethylbenzene		76	
			xylene		83	
D. TO	and and	visible	benzene	0.1	3	[01]
Fe-TiO <sub>2</sub>	sol–gel	visible		0.1 ppm		[31]
			toluene	0.1 ppm	32	
			ethylbenzene	0.1 ppm	51	
			xylene	0.1 ppm	72	50.03
N-TiO <sub>2</sub>	sol_gel	UV	ethylbenzene	1000 ppm	20	[32]
N-TiO <sub>2</sub>	sol_gel	visible	ethylbenzene	1000 ppm	13.5	[32]
N-TiO <sub>2</sub>	sol-gel	visible	toluene	-	45	[33]
N-TiO <sub>2</sub>	calcination	visible	benzene	-	72	[34]
CQDs/TiO <sub>2</sub>	impregnation	UV/visible	o-xylene	50 ppm	87	[35]
C-TiO <sub>2</sub>	sol-gel	UV	toluene	100 ppm	94	[36]
S-TiO <sub>2</sub> -AS	sol-gel	visible	benzene	1 ppm	34	[37]
S-TiO <sub>2</sub> -AS	sol-gel	visible	toluene	1 ppm	78	[37]
S-TiO <sub>2</sub> -AS	sol-gel	visible	ethylbenzene	1 ppm	91	[37]
S-TiO <sub>2</sub> -AS	sol-gel	visible	xylene	1 ppm	94	[37]
F-TiO <sub>2</sub>	hydrothermal	UV	toluene	30 ppm	80	[38]
MnO <sub>2</sub> /TiO <sub>2</sub>	hydrothermal	VUV	toluene	30 ppm	96	[39]
Fe <sub>2</sub> O <sub>3</sub> /TiO <sub>2</sub>	impregnation	fluorescent	o-xylene	50 ppm	91	[40]
CuO/TiO <sub>2</sub>	impregnation	UV + Vis	o-xylene	60 ppm	35.67	[41]
CuO/TiO <sub>2</sub>	impregnation	UV + Vis	m-xylene	60 ppm	34.48	[41]
CuO/TiO <sub>2</sub>	impregnation	UV + Vis	p-xylene	60 ppm	30.38	[41]
NO <sub>3</sub> /TiO <sub>2</sub>	hydrothermal	visible	toluene	368 mg/m <sup>3</sup>	85.3	[42]
NO <sub>3</sub> /TiO <sub>2</sub>	electrochemical	visible	toluene	200 ppm	90	[43]
ZnO/P25	impregnation	UV	toluene	500 ppm	~100	[44]
ΓiO <sub>2</sub> -CeO <sub>2</sub>	sol-gel	UV	toluene	10 ppm	96	[45]
MoS <sub>2</sub> /TiO <sub>2</sub>	hydrothermal	Visible	toluene	50 ppm	56	[46]
CuS/TiO <sub>2</sub>	SILAR	solar light	toluene	4215 ppm	40	[47]
CdS/TiO <sub>2</sub>	SILAR	solar light	toluene	4215 ppm	50	[47]
CdS/TiO <sub>2</sub>	hybridization	visible	benzene	20 μg/mL	95	[48]
g-C <sub>3</sub> N <sub>4</sub> /TiO <sub>2</sub>	impregnation	UV	toluene	200 ppm	67	[49]
$LaVO_4/TiO_2$		visible	benzene	**	70	[49]
Lav04/1102	sol–gel	visible	Denzene	250 ppm	70	[20]

<sup>a</sup> : The removal efficiency is defined as RE.

TiO<sub>2</sub> and strengthen the response of TiO<sub>2</sub> to the visible region, as shown in Fig. 1b. Metal ions, such as  $\text{Fe}^{3+}$  and  $\text{Er}^{3+}$ , dispersed in TiO<sub>2</sub> lattice break up the original lattice structure, forming lattice defects and shallow potential well [62,63]. The shallow potential well promotes the photocatalytic activity of the catalyst, which can prevent electron hole recombination. The TiO2 modified with different metals has exhibited different degradation efficiencies (Fig. 1c). The mono-modified (Pt, Cu, Ag) TiO<sub>2</sub> photocatalysts were prepared by chemical reduction method [64]. The absorption spectra of obtained photocatalysts indicated that the visible light absorption of all metal-modified photocatalysts was significantly enhanced compared to the pure titanium (IV) oxide. The types and amounts of doped metal are the decisive parameters of TiO<sub>2</sub> photocatalysis. Huang et al. [51] studied the catalytic oxidation of benzene on TiO2 modified with base metals (Mn, Co, Cu, Ni, and Fe). The presence of base metals probably promoted the transformation from anatase phase to rutile one. The shifts of the absorption on various metals were different, so did the removal efficiency of benzene (Fig. 1c). Fig. 1d shows the photodegradation of toluene by metal-TiO<sub>2</sub> with different loadings. With the amount of doped metal increasing, the degradation efficiency of toluene initially increased and subsequently decreased for all catalysts. If a certain amount of metal dopant is maintained, dopants trap electrons generated by the photoelectricity and separate charge carriers, like an electron acceptor [54,65]. However, in high dopant concentration, dopants may act as a recombination center preventing hole formation and even distort the catalyst crystallinity, which reduces the photocatalytic reactivity greatly [62,64]. As we can see in the Table 1, P25 catalysts modified with different metals were used to degrade BTEX in the gas phase under UV irradiation, and Ce-P25 exhibited better performance for BTEX degradation [29]. Additionally, the oxidation rate of the BTEX (R<sub>A</sub>, ppmv/min), the adsorption constant (K<sub>A</sub>,  $ppmv^{-1}$ ), and the rate constant (k<sub>R</sub>, ppmv/min) were calculated using the experimental data. The KA about ethylbenzene and xylene were substantially higher than those of benzene and toluene, which indicated that ethylbenzene and xylene were more easily adsorbed than benzene and toluene. The  $k_{\text{R}}$  about benzene and ethylbenzene which data obtained over Ce-P25 were higher than those about toluene and

xylene which data obtained over Ce-P25 and those about BTEX which data obtained over P25 and Fe-P25, revealing the stronger photodegradation of benzene and ethylbenzene using Ce-P25.

### 2.1.2. Doping with non-metals

The drawback of doping metals is easy to photo-corrode during the photocatalytic process. Non-metallic elements have the advantages of wide source of raw materials, low prices, and resistance to corrosion. As a result, the preparation of TiO<sub>2</sub> photocatalyst by using non-metals is a feasible way to reduce the band gap and enhance the response of visible light. Carbon (C), nitrogen (N), fluorine (F) and sulfur (S) are common TiO<sub>2</sub> non-metallic dopants used to remove VOCs (Fig. 2a) [66,67]. Doping carbon can not only increase the crystallinity of TiO<sub>2</sub>, but also make the absorption wavelength red-shift [68]. Mahmood' group also expounded that the carbon quantum dots (CQDs) notably improved the adsorption capability of TiO<sub>2</sub>, attributing to the hexagonal ring structures in CQDs [35]. Nitrogen has a low ionization energy, good stability. Doping nitrogen also can reduce the band gap of TiO<sub>2</sub>. Fig. 2b shows that the different structures of N-TiO2, including substitutional and interstitial. Li's group investigated the degradation of benzene on N-H-TiO<sub>2</sub> photocatalyst [69]. It was found that the new doping states above the VB edge and defect states appeared with N-H-codoping (Fig. 2c), which increased the surface reactive sites and reduced the recombination of charge carriers, thus improving the photocatalytic reaction. Valentin et al. [70] found that TiO<sub>2</sub> doped with carbon or nitrogen would produce alternative and interstitial crystal structures. X-ray photoelectron spectroscopy (XPS) analysis revealed that the peak at 399.6 eV was regarded as the substitutional N and the peak at 401.6 eV was attributed to the interstitial N [71]. The atomic radius of carbon and nitrogen is similar to that of the oxygen atom, and the electronegativity is weaker than oxygen. Therefore, Carbon and nitrogen may supersede oxygen in the TiO2 lattice. On the other hand, since carbon and nitrogen atoms are highly small, they may diffuse through the lattice voids and then combine with the lattice atoms. Conversely, fluorine has a higher electronegativity than oxygen. And the F 2p states are under the bottom of the O 2p VB.

The introduction of fluorine may bring about the appearance of the impurity states in the band gap of  $TiO_2$  and formation of  $Ti^{3+}$  states (Fig. 2a).  $Ti^{3+}$  states can effectively capture photo-generated electrons, which can suppress the recombination of electron-hole pairs and improve its photocatalytic performance [38]. In term of S with various valence states, the mixing of the 3p orbit of S and the valence band can reduce the forbidden bandwidth. Doping S can also facilitate the separation of electron-hole pairs [72]. S-TiO<sub>2</sub> films immobilized on aluminum sheets exhibited excellent performance for decomposing the BTEX under visible light irradiation [37].

# 2.1.3. Doping with semiconductors

The composites of semiconductors are catalysts obtained by compounding different semiconductors with titanium dioxide nanomaterials by using the coupling effect between nanoparticles. Composites outperform single semiconductors because the difference of chemical potential between semiconductors causes band bending at the interface of junction. A built-in field is generated by the band bending, resulting in a spatial separation of the electrons and holes on different sides of heterojunction. There are several semiconductors with different band edges, and semiconductor compounding includes three methods (Fig. 3a). The first is to connect with narrow-bandgap semiconductors (CdS, C<sub>3</sub>N<sub>4</sub>, etc.). For instance, combined with CuS and CdS, the composites of CuS-CdS/TiO<sub>2</sub> generated a stepwise energy level structure because the CB and VB edges of the three materials both increased in the order of  $TiO_2 < CdS < CuS$ , thereby obtaining a higher driving force for the electron injection at the interfaces and highly suppressing the recombination of electron-hole pairs [47]. The second is to use wider band metal oxides (ZnO, SnO<sub>2</sub>, etc.) to suppress electron recombination. The heterojunction was formed between  $TiO_2$  and  $SnO_2$  (Eg = 3.8 eV), and the electrons generated from  $TiO_2$  were shifted to  $SnO_2$  [74]. The charge-transfer process caused a better performance. The third is to combine with p-type semiconductor (CuO, etc.) on n-type TiO<sub>2</sub> semiconductor to improve the photocatalytic efficiency. The energy band structure and density of TiO2 and CuO-TiO2 composites were calculated



**Fig. 2.** (a) Schematic energy level of  $TiO_2$  doped with non-metals. (b) The structures of the three different forms of *N*-dopped Anatase [73]. (c) Schematic diagram of energy states and photocatalytic reaction paths on N–H– $TiO_2$  nanoparticles [69].



Fig. 3. (a) Band edge position of diverse semiconductors, using the normal hydrogen electrode (NHE) as a reference [75]. (b) Band structure and (c) density of state of TiO<sub>2</sub> and CuO-TiO<sub>2</sub> composites [41]. (d) The photocatalytic mechanism of Pd-Ag/AgBr/ TiO<sub>2</sub> hierarchical nanostructure [76]. (e) Z-scheme charge-transfer mechanism for the  $\alpha$ -MnO<sub>2</sub>/Mn<sub>3</sub>O<sub>4</sub> system [77].

by the density functional theory (DFT) [41]. As shown in Fig. 3b, the indirect band gap of  $TiO_2$  was 2.84 eV initially, and the band gap reduced to 2.71 eV and even to 0.14 eV with the introduction of CuO. Fig. 3c displays that the orbit of Cu 3d and O 2p participated in the hybridization and generated more impure states, thereby leading to a lower band gap.

Sulfides such as MoS<sub>2</sub> (1.75 eV), CuS (2.00 eV), and CdS (2.40 eV) are common narrow-bandgap semiconductors. The CuS/TiO<sub>2</sub> and CdS/TiO<sub>2</sub> photocatalysts were successfully prepared via simple successive ionic layer adsorption and reaction (SILAR) method used for removal of toluene [47]. The absorption edge of  $TiO_2$  had a red shift on account of the loading of CuS or CdS composition, resulting in a better visible light response ability. The PL emission spectra elucidated that the order of the recombination of free carriers was as follows CuS/TiO<sub>2</sub> nanobelts < CdS/TiO<sub>2</sub> nanobelts < TiO<sub>2</sub> nanobelts. In addition, MoS<sub>2</sub> could enhance the photocatalytic activity with the exposed sulfur atoms behaving as active sites. Hence, MoS<sub>2</sub>/TiO<sub>2</sub> composites displayed a predominant degradation efficiency of gaseous toluene, prepared by a facile hydrothermal reaction [46]. A series of  $TiO_2/MnO_2$  nanosheets with different percentage of Mn were characterized by UV-vis diffuse reflectance spectra (DRS). The absorption range of samples was extended from 410 nm to near infrared light. Besides that, as the amount of MnO<sub>2</sub> increased, the light absorption intensity of TiO<sub>2</sub>/MnO<sub>2</sub> nanosheets in visible light regions increased [39]. Nevertheless, ZnO/P25 composites prepared by an incipient wetness impregnation method degraded the gas toluene exposing to the UV lamp [44]. With the increase of ZnO content, the light absorption intensity increased at the initial stage and then decreased, and the toluene conversion rate also increased firstly and then decreased. There was a special charge transfer mechanism of the bimetallic Pd-Ag co-catalyst on AgBr/TiO<sub>2</sub> composites in Fig. 3d [76]. The AgBr/TiO<sub>2</sub> heterostructure was a type-II semiconductor, which enhanced the charge separation and the visible light utilization. In addition, the existence of Ag and Pd improved the charge transfer, thereby greatly promoting the photocatalytic process.

## 2.2. Non-TiO<sub>2</sub> photocatalysts

The major types of non-TiO<sub>2</sub> photocatalysts used on BTEX removal are monadic or binary metal oxides, hydroxides, Bismuth-based materials, and metal-free materials. Those materials with different bandgap obtained different methods were listed in Table 2, which were used in the photodegradation of BTEX.

#### 2.2.1. Metal-oxide-based semiconductor

There are many valence states for manganese. Different valence states of manganese oxides can form heterojunctions, which can effectively prevent the recombination of photogenerated electrons and holes, thus improving the photocatalytic activity. The hierarchical porous  $\alpha$ -MnO<sub>2</sub>/Mn<sub>3</sub>O<sub>4</sub> nanostructures with multiple interfaces were successfully designed and synthesized [77]. As shown in the Fig. 3e, the direct

#### Table 2

The photodegradation of BTEX using non-TiO<sub>2</sub> photocatalysts.

Photocatalyst	Light	BETX	Concentration	RE (%)	Reference
ZnO	sunlight	benzene	50 ppm	57	[78]
ZnO	sunlight	toluene	50 ppm	67	[79]
ND-ZnO	UV	toluene	50 ppm	100	[80]
Si-GO/ZnO	UV	toluene	100 ppm	76.18	[81]
Silica-GO/ZnO	UV	benzene	50 mg/m <sup>3</sup>	87.20	[82]
ZnO-13X zeolite	UV	benzene	5 ppm	98.4	[83]
ZnO-AC	UV	benzene	5 ppm	99.0	[83]
E-MnO <sub>2</sub> /C	visible	toluene	56 mmol/L	87.34	[84]
α-MnO <sub>2</sub> /Mn <sub>3</sub> O <sub>4</sub>	Xenon lamp	toluene	300 ppm	>90	[77]
M-CeO <sub>2</sub>	VUV	toluene	30 ppm	82.0	[85]
N-CeO <sub>2</sub>	VUV	toluene	30 ppm	87.8	[85]
T-CeO <sub>2</sub>	VUV	toluene	30 ppm	87.9	[85]
WO <sub>3</sub>	Xenon lamp	toluene	$2800 \text{ mg/m}^3$	73.8	[86]
ZnCr–LDHs	visible	o–xylene	$175 \text{ mg/m}^3$	40.9	[87]
	visible	-		83.6	
Au/ZnCr-LDHs		o-xylene	$175 \text{ mg/m}^3$		[87]
Au/ZnCr–LDHs	visible	m–xylene	$175 \text{ mg/m}^3$	79.7	[87]
Au/ZnCr–LDHs	visible	p–xylene	175 mg/m <sup>3</sup>	77.8	[87]
Au/ZnCr–LDHs	visible	benzene	$175 \text{ mg/m}^3$	97.7	[87]
Au/ZnCr–LDHs	visible	toluene	$175 \text{ mg/m}^3$	91.4	[87]
MgSn(OH) <sub>6</sub>	UV	benzene	250 ppm	~70	[88]
ZnSn(OH) <sub>6</sub>	UV	benzene	250 ppm	88	[88]
CaTiO <sub>3</sub>	UV	toluene	50 mg/L	~65	[89]
Eu <sup>3+</sup> -CaTiO <sub>3</sub>	UV	toluene	50 mg/L	~80	[89]
Pt/SrTiO <sub>3</sub>	visible	toluene	700 ppm	85	[90]
BaZrO <sub>3</sub>	visible	toluene	_	~34	[91]
SrZrO <sub>3</sub>	visible	toluene	_	~23	[91]
LaFeO <sub>3</sub>	UV-vis	p-xylene	19 mg/L	35.8	[92]
LaMnO <sub>3</sub>	UV–vis	p-xylene	19 mg/L	37.6	[92]
LaCoO <sub>3</sub>	UV–vis	p-xylene	19 mg/L	26.1	[92]
LaFeO <sub>3</sub>	visible	benzene	_	85	[93]
KTaO <sub>3</sub>	visible	toluene	150 ppm	47	[94]
KTaO <sub>3</sub>	visible	toluene	200 ppm	13.11	[95]
Pt/Er-KTaO <sub>3</sub>	visible	toluene	200 ppm	30.92	[95]
Rh/Pr-KTaO <sub>3</sub>	visible	toluene	200 ppm	34.73	[95]
	visible	toluene		17.13	
KTaO <sub>3</sub>			200 ppm		[96]
Au-KTaO3	visible	toluene	200 ppm	29.29	[96]
Ag-KTaO <sub>3</sub>	visible	toluene	200 ppm	30.43	[96]
Pt-KTaO <sub>3</sub>	visible	toluene	200 ppm	36.42	[96]
Pd-KTaO <sub>3</sub>	visible	toluene	200 ppm	33.34	[96]
Rh-KTaO <sub>3</sub>	visible	toluene	200 ppm	41.98	[96]
Ru-KTaO <sub>3</sub>	visible	toluene	200 ppm	25.68	[96]
ZnFe <sub>2</sub> O <sub>4</sub>	UV	toluene	900 ppm	~60	[97]
ZnFe <sub>2</sub> O <sub>4</sub>	visible	toluene	900 ppm	~34	[97]
CaFe <sub>2</sub> O <sub>4</sub>	visible	toluene	-	76	[98]
ZnGa <sub>2</sub> O <sub>4</sub>	UV	benzene	220 ppm	20	[99]
ZnGa <sub>2</sub> O <sub>4</sub>	UV	toluene	180 ppm	26	[99]
ZnGa <sub>2</sub> O <sub>4</sub>	UV	ethylbenzene	190 ppm	30	[99]
ZnCo <sub>2</sub> O <sub>4</sub>	UV	toluene	-	78	[100]
Bi <sub>2</sub> WO <sub>6</sub>	visible	toluene	250 ppm	80.6	[101]
Pt/Bi <sub>2</sub> WO <sub>6</sub>	visible	toluene	250 ppm	>90	[101]
Bi <sub>2</sub> WO <sub>6</sub>	visible	toluene	$350 \text{ mg/m}^3$	45.9	[102]
CQDs/Bi <sub>2</sub> WO <sub>6</sub>	visible	toluene	$350 \text{ mg/m}^3$	96.9	[102]
Bi <sub>2</sub> WO <sub>6</sub>	visible	benzene	100 ppm	~38	[102]
BiOCl/ Bi <sub>2</sub> WO <sub>6</sub>	visible	benzene	100 ppm	~20	[103]
BiOCI/ BI2WO6 BiOCI	visible	benzene	100 ppm	$\sim 20$ $\sim 12$	[103]
BiOCI	UV				
		toluene	50 ppm	52.5	[104]
BiVO4	visible	toluene	25 ppm	~30	[105]
BiO <sub>1.2</sub> I <sub>0.6</sub>	Xenon lamp	toluene	$2800 \text{ mg/m}^3$	82.3	[86]
BiO <sub>1.2</sub> I <sub>0.6</sub> /WO <sub>3</sub>	Xenon lamp	toluene	2800 mg/m <sup>3</sup>	>95	[86]
(BiO) <sub>2</sub> CO <sub>3</sub>	visible	toluene	1200 ppm	~22.5	[106]

Z-scheme heterojunction mechanism indicated that the photogenerated electrons in the CB of  $\alpha$ -MnO<sub>2</sub> migrated to the VB of Mn<sub>3</sub>O<sub>4</sub> and recombined with holes, leading to effective spatial charge separation. The accumulated electrons in the CB of Mn<sub>3</sub>O<sub>4</sub> were captured by the absorbed oxygen on the surface to generate  $\bullet$ O<sub>2</sub><sup>-</sup> radicals that participated in photocatalysis. MnO<sub>2</sub> with a narrow band gap, has a wide range of absorption under sunlight. Zhou *et al.* [84] discovered a 3D hierarchical  $\epsilon$ -MnO<sub>2</sub> pasted with carbon. The C-coated  $\epsilon$ -MnO<sub>2</sub> could expand the photo-response from UV to visible light region, thus generating more electron-hole pairs and exhibiting unexceptionable performance for removing toluene.

ZnO is a promising semiconductor material which has a broad band gap width (3.2 eV, Fig. 3a) and a large excitation binding energy (60 meV) [107]. Furthermore, the absorption efficiency of ZnO in the solar spectrum is higher than TiO<sub>2</sub>. ZnO is considered an attractive alternative to titanium dioxide on account of their similar chemical-physical characteristics. ZnO also has the advantages of stable chemical properties, abundant appearance, biological safety and low cost. The catalytic degradation of benzene and toluene by ZnO coated on glass plates was studied under simulated sunlight, which had a relatively superior efficiency [78,79]. Like TiO<sub>2</sub>, doping ZnO with metal or nonmetal displayed a highly efficient degradation of BTEX, as seen in Table 2. The noble metal can effectively capture electrons as an electron acceptor [108]. Modified with Ag, ZnO had excellent activity for benzene removal on account of the strong separation of photo-generated electrons and holes [108]. The Pd was introduced into the ZnO superstructures doped by carbon using the sol- immobilization process, which had better photothermal catalytic performance towards benzene oxidation [109]. The existence of defects and oxygen vacancies caused by doping C, promoted the oxygen species transportability and enhanced the catalytic activity as active sites [110]. In addition, some carbon materials are also commonly used to improve photocatalytic performance. Graphene can prevent the recombination of charges during electron transport, thereby improving the photocatalytic activity of ZnO. On the other hand, the high specific surface area and high adsorption rate of graphene enhance the contact between pollutants and ZnO, which promotes pollutants degradation [81,82]. However, the photocorrosion of ZnO is a common and serious phenomenon that often occurs in the long-term photocatalysis process, which greatly hinders its further application. Liu et al. [80] successfully developed a strategy to suppress photocorrosion of ZnO with introducing the nanodiamonds (ND). The modification of ND not only effectively inhibited the photocorrosion to prolong the working life, but also greatly enhanced the photocatalytic ability. With the decoration of ND, free electrons of ND could transfer to ZnO to form a built-in electric field, and the photo-generated  $h^+$  and formed  $O_2^-$  participated in the degradation of toluene (Fig. 4a).

CeO<sub>2</sub> is a rare-earth oxide that has a broad direct band gap width (3.2 eV) and light absorption edge (388 nm) [16]. CeO<sub>2</sub> has excellent photocorrosion resistance, and has excellent ultraviolet light absorption capacity. However, it is more important that the active surface sites can arise by switching readily between Ce<sup>4+</sup> and Ce<sup>3+</sup>, giving rise to the formation or destruction of associated oxygen vacancy easily. And the redox properties of Ce<sup>4+</sup>/Ce<sup>3+</sup> pairs might lead to the motivation of hydroxyl radical generation [111]. Also, there are some correlations between the activity and structure of CeO<sub>2</sub>. Mesoporous microcuboid CeO<sub>2</sub> (M–CeO<sub>2</sub>), CeO<sub>2</sub> nanorods (N-CeO<sub>2</sub>), and CeO<sub>2</sub> with tunable porosity (T-CeO<sub>2</sub>) were formed and applied for the removal of toluene under VUV irradiation [85]. The UV–vis DRS demonstrated that three



Fig. 4. (a) The proposed mechanism for photocatalytic degradation of toluene upon nanodiamond-decorated ZnO [80]. (b) The structures of BiOCl and  $Bi_2WO_6$  in the Bi-W-Cl-O solvothermal system [103]. (c) The novel hollow BiOCl@CeO<sub>2</sub> heterostructure [114]. (d) The proposed heterojunction mechanisms between  $\alpha$ -Fe<sub>2</sub>O<sub>3</sub> and g-C<sub>3</sub>N<sub>4</sub> [115].

samples had stronger absorbance in the UV region and weaker absorbance in the visible region, relatively. The results of VUV-photocatalytic oxidation indicated that N-CeO<sub>2</sub> and T-CeO<sub>2</sub> exhibited higher removal efficiency of toluene. Pd-Ag@CeO<sub>2</sub> which had a core–shell structure significantly improved the light adsorption and charge separation in the visible region [112]. The synergistic effect of Pd and Ag weakened the Ce-O bond and produced a high concentration of lattice defects, which further promoted the transformation of lattice oxygen and adsorbed oxygen.

 $WO_3$  has a broad direct band gap width (2.8 eV, Fig. 3a). And  $WO_3$  can more effectively utilize visible light, which accounts for nearly half of the solar radiation energy. Wang *et al.* [86] found that  $WO_3$  exhibited remarkable catalytic performance for photocatalytic toluene. However, pure  $WO_3$  is difficult to maintain stable photocatalytic performance, thanks to defects such as easy photocorrosion and low utilization rate of visible light. Hence,  $WO_3$  is often doped with metals or coupled with other semiconductors to improve its photocatalytic performance [43,113].

# 2.2.2. Hydroxide photocatalysts

Most degradation reactions are mediated by hydroxyl radicals (•OH), and the formation of hydroxyl radicals can be effectively promoted by surfaces with high hydroxyl density. It is necessary to generate OH radicals in the process of photocatalytic oxidation, so hydroxides are also often used as photocatalysts to degrade VOCs. A series of MSn(OH)<sub>6</sub> (M = Co, Cu, Fe, Mg, Mn, Zn) was synthesized and exhibited obvious visible absorptions [88]. Only MgSn(OH)<sub>6</sub> and ZnSn(OH)<sub>6</sub> displayed a predominant and stable activity for the destruction of benzene. In these samples, the hydroxyl structure gave them high performance, promoted the formation of active radicals ( $\bullet$ OH and  $\bullet$ O<sub>2</sub>), and inhibited the deposition of stable by-products. According to the research, ZnCr-LDHs (Layered Double Hydroxides) with narrow band bap had considerable reactivity for photocatalytic reaction, which had a good deal of water molecules and hydroxyl ions [87]. The water molecules and hydroxyl ions were the key to the generation of OH radicals during photocatalytic oxidation. Above all, the active species experiments revealed that the major drive of photocatalytic reaction was the hole and hydroxide radicals for Au supported ZnCr-LDHs. This indicated that Au/ZnCr-LDHs exhibited higher photocatalytic activity.

# 2.2.3. Perovskite oxides

Perovskite (ABO<sub>3</sub>) oxide has become one of the researches focuses in the field of photocatalysis due to its stable structure, narrow band gap, wide light absorption wavelength range, and high solar energy utilization rate. The CaTiO<sub>3</sub> perovskites were studied for photocatalytic degradation of toluene [89]. The adsorption edges of the CaTiO<sub>3</sub> perovskites shifted to a longer wavelength of ca. 400 nm. The perovskite LaMO<sub>3</sub>, as a highly efficient photocatalyst rich in earth, has attracted great attention. This structure benefits from the possibility of replacing M. Different M not only results in controlling the band gap, but also ultimately controlling the energy arrangement of the surface, such as oxygen vacancies which can contribute to the higher photocatalytic activity. The perovskite-like oxides  $LaBO_3$  (B = Fe, Co, and Mn) were prepared by sol-gel method which exhibited ascendant stability and activity for photodegradation of p-xylene under radiation of UV-vis light [92]. It indicated that the band gap energy (Eg) for perovskites LaBO<sub>3</sub> was about 1.87–2.2 eV, leading to absorption increasing in the visible region. Jamali et al. [93] found that LaFeO3 nanoparticles had a narrow band gap energy (2.12 eV) and mesoporous structure, with high crystallinity. KTaO3 was also commonly used as a photocatalyst with high photochemical stability and a relatively wide band gap of about 3.4 eV (Fig. 3a), resulting in that it can only be excited by UV. Krukowska et al. [95,96] studied that KTaO<sub>3</sub> decorated with mono or bimetallic nanoparticles had predominant Vis and UV-vis light activity for toluene removal in gas phase. Therefore, Metals are often doped in KTaO<sub>3</sub> for broadening light absorption to longer wavelengths.

#### 2.2.4. Spinel oxides

Spinel-type complex oxide (AB<sub>2</sub>O<sub>4</sub>) has advantages of visible light response, stable structure, diverse composition, narrow band gap, etc. It can adjust its structure while maintaining the original configuration unchanged and improve the visible light absorption efficiency of the materials. Hence, spinel oxide is an ideal material for exploring new photocatalysts with large visible light absorption range and high quantum conversion efficiency, as illustrated in Table 2. Zhou et al. [100] established hollow porous zinc cobaltate (ZnxCo3-xO4) nanocubes in order to heighten the photocatalytic performance of spinel oxides. Compared with traditional stoichiometric ZnCo<sub>2</sub>O<sub>4</sub> nanoparticles, the Zn<sub>x</sub>Co<sub>3-x</sub>O<sub>4</sub> nanocubes had a higher specific surface area, stronger light absorption capacity in the range of 350-800 nm, and lower recombination rate of photogenerated electron-hole pair. And Fe-base spinel materials exhibit high catalytic stability on account of the framework structure of Fe-O octahedrons. Meanwhile, The Fe d-orbitals of AFe<sub>2</sub>O<sub>4</sub> play a major role in the formation of narrow band gaps, which can broaden light absorption to longer wavelengths. Using spinel-type mesoporous nano-ZnFe<sub>2</sub>O<sub>4</sub> (1.87 eV) as photocatalyst, toluene, a metabolite of VOCs pollutants, can be effectively removed under UV and visible light irradiation [97]. Yin et al. [98] had designed a novel hedgehog-like CaFe<sub>2</sub>O<sub>4</sub> hollow-structured materials by a simple template-assisted method. The characterization results showed that this new structure could significantly improve the separation and transformation of photogenerated  $e^{-}/h^{+}$  pairs.

## 2.2.5. Bismuth-based catalysts

Bismuth-based compounds have become a research hotspot in the field of non-TiO<sub>2</sub> photocatalyst development in recent years due to their special layered structure and appropriate band gap. From the perspective of its electronic structure, the hybridization of 6s orbital of Bi and 2p orbital of O causes its valence band to move up, which is conducive to the separation and migration of photo-generated  $e^-$ - $h^+$  pairs during photocatalysis. Therefore, bismuth-based photocatalysts have good photocatalysts commonly used to catalyze VOCs include ternary metal oxides, bismuthate and bismuth oxyhalides.

A typical binary metal oxide Bi2WO6 has an orthorhombic Aurivillius-type structure composed of alternating bismuth oxide  $[Bi_2O_2]^{2+}$  and octahedral  $(WO_6)^{2-}$  (Fig. 4b). Due to the narrow band gap (2.5–2.8 eV), Bi<sub>2</sub>WO<sub>6</sub> has strong absorption under visible light and favorable light stability. The Bi<sub>2</sub>WO<sub>6</sub> composites modified with Pt have a high concentration of oxygen vacancies [101]. The oxygen vacancies below the CB might reduce the band gap of Bi<sub>2</sub>WO<sub>6</sub> and facilitate the formation of  $\bullet O_2^-$ , accelerating the mineralization of gaseous toluene. And simultaneously, the deposition of Pt promoted the separation of photogenerated charge pairs. Liu et al. [102] introduced carbon quantum dots (CQDs) in Bi<sub>2</sub>WO<sub>6</sub> to remove toluene under visible light, which makes the light absorption red-shifted and oxidizing ability enhanced. Furthermore, the CQDs/Bi<sub>2</sub>WO<sub>6</sub> behaved high stability and reusability. And monoclinic BiVO<sub>4</sub> is another common catalyst with high activity. The estimated band gap of BiVO<sub>4</sub> is 2.17 eV giving rise to strong absorption of visible light [105].

Bismuth oxyhalides BiOX (X = F, Cl, Br, I) have outstanding optical and electrical properties, which are crystallized in a tetragonal matlockite structure. As shown in Fig. 4b, BiOCl is a layered semiconductor photocatalyst composed of  $[Bi_2O_2]^{2+}$  layers and Cl<sup>-</sup> alternately combined. BiOCl with a large band gap (3.4 eV) has high insensitivity to visible light [103]. Three stages including BiOCl, BiOCl/Bi<sub>2</sub>WO<sub>6</sub> composites, and Bi<sub>2</sub>WO<sub>6</sub> were prepared by changing the reaction conditions in the Bi-W-Cl-O system [103]. The UV–vis DRS spectra revealed that Bi<sub>2</sub>WO<sub>6</sub> had a higher light efficiency than BiOCl. In addition, the photocatalytic efficiency of the three samples for benzene under visible light was arranged as Bi<sub>2</sub>WO<sub>6</sub> > BiOCl/Bi<sub>2</sub>WO<sub>6</sub> composites > BiOCl. Dong *et al.* [104] successfully constructed the surface oxygen vacancies (OVs) on the surface of BiOCl via the solvothermal-induced hot ethylene glycol reduction. The introduction of OVs enhanced the adsorption and activation of O2, H2O and toluene molecules to generate more reactive free radicals with higher oxidation capacity. As we can see in Fig. 4c, a type-II heterostructure was formed between p-type BiOCl and n-type CeO<sub>2</sub>, which established a new Fermi level balance leading to the alteration of energy band of two semiconductors [114]. The internal electric field was built by the transfer cycle of  $h^+/e^-$ , and promoted the separation of  $h^+/e^$ e<sup>-</sup>, thus improving the photocatalytic performance. BiOI is an attractive semiconductor with a layered crystal structure and a suitable band gap. Nonetheless, the rapid recombination of charge and poor electrical conductivity lead to low photocatalytic efficiency, which hinders the practical application of BiOI. Herein, Wang et al. [86] synthesized a novel photocatalytic composite material of I-deficient BiO1.2I0.6/WO3 by solvothermal method coupling with thermal decomposition. The introduction of iodine vacancy and WO<sub>3</sub> prominently strengthened the photocatalytic activity of BiOI.

 $(BiO)_2CO_3$  is an orthogonal crystal structure in which  $[Bi_2O_2]^{2+}$  layer and  $[CO_3]^{2-}$  layer are alternately arranged. The pure  $(BiO)_2CO_3$  can only absorb the UV light up to 370 nm, which can be assigned to the band gap of  $(BiO)_2CO_3$  for 3.2 eV. To expand the light absorption of  $(BiO)_2CO_3$ , Duan *et al.* [106] successfully synthesized a ternary heterojunction  $(BiO)_2CO_3$ -CdS-CuS. It was found that the band gap of the  $(BiO)_2CO_3$ -CdS-CuS heterojunction was reduced to about 2.25 eV, which showed preeminent catalytic activity and stability for toluene degradation.

# 2.2.6. Graphitic carbon nitride

The simple graphitic carbon nitride (g-C<sub>3</sub>N<sub>4</sub>) with a suitable band gap is an ideal candidate material for a new generation of the visible light responsive photocatalyst. The g-C<sub>3</sub>N<sub>4</sub> is a non-toxic, low-cost, metal-free, stable semiconductor made from readily available starting materials containing the most abundant elements (carbon and nitrogen). Zhao et al. [116] studied that the catalytic activity of pure g-C<sub>3</sub>N<sub>4</sub> for benzene was weaker than that of TiO2. The low specific surface area and high photocarrier recombination rate limited the application of g-C<sub>3</sub>N<sub>4</sub>. Therefore, g-C<sub>3</sub>N<sub>4</sub> is often combined with other semiconductors to enhance its activity. The custom-designed H3PW12O40/g-C3N4 filmcoated optical fiber photoreactor was demonstrated for removal of aromatics including benzene, toluene and m-xylene under simulated sunlight [117]. The H<sub>3</sub>PW<sub>12</sub>O<sub>40</sub>/g-C<sub>3</sub>N<sub>4</sub> film exhibited remarkably higher photocatalytic activities which had better stability and recyclability. A novel α-Fe<sub>2</sub>O<sub>3</sub>/Cu/g-C<sub>3</sub>N<sub>4</sub> nanocomposite successfully prepared by a simple and facile procedure showed excellent photocatalytic performance and stability for the removal of air ethylbenzene [115]. As shown in Fig. 4b, there were two probable mechanisms including type-II heterojunction and Z-scheme system. The type-II heterojunction mechanism can't produce ·OH and ·O2 radicals due to the CB edge potential of  $\alpha$ -Fe<sub>2</sub>O<sub>3</sub> and the VB edge potential of g-C<sub>3</sub>N<sub>4</sub>. The Z-scheme mechanism was rational in which the fast combination happened between the CBelectrons of  $\alpha$ -Fe<sub>2</sub>O<sub>3</sub> and the VB-holes of g-C<sub>3</sub>N<sub>4</sub>. Furthermore, the introduction of Cu promoted the transformation and combination of weak charge carriers.

# 2.3. Summary

In conclusion, it is important for the photocatalytic materials to convert light energy into chemical energy in the photocatalytic degradation of BTEX, which could generate the electron-hole to form the active oxygen species. Although, semiconductors with a narrow band gap excites electrons to advance from the valence band to the conduction band with relatively low photon energy, the lifetime of photogenerated electron is too short. Through investigation, it is found that both TiO<sub>2</sub>-based and non-TiO<sub>2</sub> photocatalytic materials exhibited different performance on BTEX removal. However, compared with semiconductor monomers or doping elements, heterojunction composites show excellent performance, so it is worth further study on forming the heterojunction using the semiconductors with different bandgap. There is a band bending at the interface of heterojunction owing to the difference of valance and conduction bands, and a built-in field might be formed. The built-in field could drive the electrons and holes to move in opposite directions, which lead lower recombination rate of photocarrier to generate more active oxygen species. Therefore, it is crucial to choose the semiconductors and design the heterojunction for photodegradation of BTEX.

# 3. The effect of operating parameters

## 3.1. Operation mode of photocatalytic system

The mode of operation plays a vital role in the photocatalytic system of gaseous BTEX, which may dramatically alter performance. Generally speaking, there are two modes of photocatalytic operation including continuous mode and batch mode. For continuous mode system, a certain concentration of gaseous BTEX continuously is carried out in reactor. Due to the limited exposure time and sustained high concentration, the continuous mode system exhibits lower performance at the same light source [31,41]. The continuous mode system is mostly used with the smaller footprint and larger handling capacity in large-scale industries [118]. On the contrary, the batch mode system includes a closed reactor. The target pollutants could interact with active species and be degraded deeply in the closed reactor, which exhibits higher performance [80,119]. The batch mode usually applies to some smallscale industries owing to its simple operation and maintenance [120].

#### 3.2. Initial concentration and flow rate

Photocatalytic oxidation which is heterogeneous reaction generally includes three processes: adsorption, surface reaction, and desorption. For the surface reaction, active species oxidize a certain amount of BTEX into small molecules and even mineralize to H<sub>2</sub>O and CO<sub>2</sub>. On the one hand, increasing the concentration of BTEX drives the kinetics [97]. And a higher initial concentration could boost the probability of active substances or high-energy particles colliding with BTEX molecules [121]. On the other hand, the active species are not enough to oxidize all BTEX as the concentration of BTEX increased. Furthermore, more intermediate products and by-products are produced with the initial contaminant concentration rising. The generation of intermediate products and by-products inhibits the mineralization and enhances the competitive adsorption [47]. Fu et al. [87] found that the removal rate felled slowly when the initial concentration was from 43.6 to 175 mg/  $m^3$ . When the concentration of o-xylene was at 218 mg/m<sup>3</sup>, the removal rate felled tremendously. The initial toluene concentrations were studied from 300 to 3200 ppm demonstrating that the RE first increased and then decreased with the increase of toluene concentration [97]. More of the adsorbed toluene blocked the formation of hydroxyl radicals on the photocatalyst surface. The effect of the initial concentration of BTEX on photodegradation was demonstrated [122]. As we can see in Fig. 5, when the concentration of BTEX increased from 1 ppm to 5 ppm, the degradation efficiency decreased successively. The reason for this change might be the competition of pollutants and the inhibitory effect of intermediate products or by-products [22].

The flow rate has different effects on the photocatalytic reaction in the continuous mode of operation. The residence time of gas decreases with the increase of airflow rate. Too short residence time will result in low photodegradation efficiency, considering that BTEX gases spend too little time on the catalyst and BTEX does not react adequately [123]. The lower flow rate also reduces convection and diffusion in photocatalytic reactors. On the contrary, too long residence time will lead to the production of a large number of by-products and shorten the service life of the catalyst. Jafari *et al.* [124] discussed the effect of flow rate on toluene decomposition and found that the significantly reduced RE could be attributed to the short contact time. On the other hand, the mass transfer



Fig. 5. The photocatalytic degradation efficiencies (PDE) of BTEX using g-C<sub>3</sub>N<sub>4</sub>/Cu/TNT composite under UV irradiation at different inlet concentrations (a), air flow rates (b), and relative humidity (c) [122].

rate between pollutant and catalyst increases with the gas flow rate increasing. Moreover, the mass transfer promotes the photocatalytic oxidation processes. As the mass transfer rate increase, the removal of pollutants significantly improved [125]. Shayegan et al. [126] found that the mass transfer emerges was regarded as the controlling parameter, resulting in the reduction of removal efficiency tinily at the airflow rates between 0.015 and 0.020 m<sup>3</sup>/min, respectively. Both the RE and the reaction rates were studied over P25 to investigate the effect of flow rate. When the air flow rate increased from 0.2 to 2 L/min and other operating conditions remain unchanged (concentration of toluene was 3.8 mg/m<sup>3</sup>, light intensity was 95 W/m<sup>2</sup>, and relative humidity was 25 %), the RE decreased to 10 % of the initial efficiency and the reaction rates were almost constant, which revealed that the surface photocatalytic reactions were the controlling process rather than the gas-solid phase mass transfer [11]. In a word, mass transfer is the main factor in controlling the reaction rate at lower flow rate. However, the effect of residence time was more remarkable in controlling the reaction when the flow rate was higher [47].

#### 3.3. Temperature and relative humidity (RH)

The reaction temperature plays an important role in the photocatalytic process. The higher temperature inhibits the adsorption of pollutants because the adsorption process is usually exothermic, and has a positive impact on the desorption process of intermediate and byproducts [78]. In addition, the rise of temperature accelerates the photocatalytic oxidation kinetic reaction according to the Arrhenius formula [79]. Xin et al. [47] investigated the toluene photodegradation at different temperatures. And the CuS-CdS/TO2 showed the best removal efficiency at the reaction temperature of 40°C, which might attribute to the different rate-determining steps at different reaction temperatures. Furthermore, combining thermo-catalysis with photocatalysis to maximize synergistic effects has been widely used for VOCs removal [127,128]. Both the thermal and photo excitation facilitate the effective reaction between reactive oxygen species and BTEX molecules. Yang et al. [90] studied that the photothermocatalytic oxidation of toluene was much larger than the single thermo-catalytic or photocatalytic process, suggesting the generation of photo-thermal synergistic catalytic property.

Water molecules are widely present in the atmosphere, especially in all kinds of industrial emissions. In general, the presence of water molecules has a double effect on the photocatalytic oxidation process of BTEX·H<sub>2</sub>O as a reactant can generate reactant radical species playing an important part in the oxidation process [41,129]. On the other hand, H<sub>2</sub>O may compete with the BTEX and oxygen for adsorption sites on the surface of catalyst [122,130]. At higher humidity, water molecules could connect to OH· present on the photocatalyst surface via hydrogen bond resulting in the formation of multilayer film on the catalyst surface, which can effectively inhibit the photocatalytic degradation of pollutants [79]. Due to the influence of both positive and negative aspects, the degradation efficiency of BTEX showed different trends with the increase of RH. For example, the degradation efficiency increased first and then decreased with the increase of relative humidity in some studies [78,82,126]. However, other researchers found that the degradation efficiency decreased first and then increased as the relative humidity increased [97]. The highest RE of toluene using Ag/TiO<sub>2</sub>/CA catalyst was obtained at 45 % humidity, while water may also accelerate the recombination of electron-hole pairs when further increased RH levels, resulting in the inhibition of photocatalytic activity [124]. Zhou *et al.* [131] deduced that high humidity could lead to the annihilation of high-energy electrons and the reduction in the number of electrons owing to the electronegativity of water molecules.

# 3.4. Light source

The light source has an oversized impact on photocatalytic process of BTEX. First of all, it's worth noting the wavelength of light. Ultraviolet light has a wavelength range of less than 400 nm, while visible light has a wavelength range of 400–760 nm. Generally speaking, there are many photocatalysts, which are insensitive to visible light [43,63,132]. Hence, many researchers have demonstrated the RE of BTEX under UV light was higher than under the visible light [32,133]. The intensity of light also greatly affects the photocatalytic reaction. The enhancement of light intensity in photocatalytic reaction will generate more photon energy and further stimulate the generation of electron hole pairs and the formation of active substances, effectively improving the photocatalytic degradation [97,126]. In addition, BTEX photocatalytic process may produce more byproducts under low light intensity. Fu et al. [87] studied the rate of photocatalytic reaction and discovered that the stronger the light intensity, the faster the photocatalytic reaction rate. Oppositely, high light intensity has a negative effect on photocatalytic process, because the recombination rate of electron hole pairs increases with the enlargement of light intensity [124].

#### 3.5. Summary

The photocatalytic degradation of BTEX is collectively affected by different operating conditions. It is difficult to identify-one main role in the photocatalytic degradation of BTEX with the different materials and different reaction mechanisms. Liang's group found that the air flow rate and illumination intensity played the main role on BTEX removal [134]. In the practical application, the air flow rate and the light are more easily adjusted and controlled than other parameters, and it is important to explore the optimal air flow rate and light intensity.

# 4. Reaction mechanism of BTEX photocatalysis

#### 4.1. Reaction kinetics

The kinetic process of photocatalytic reaction not only determines

the degradation efficiency, but also determines the reaction rate, which is essential to the practical application. It is reported that the process of photo-oxidizing BTEX most widely follows the Langmuir-Hinshelwood (L-H) model [49,135–138]. The BTEX molecules are adsorbed on the surface of catalyst, firstly. And photo-introduced electron-hole pairs may react with O<sub>2</sub> or H<sub>2</sub>O adsorbed on the catalyst surface to form surface adsorbed reactive oxygen species. Then the surface adsorbed reactive oxygen species oxidized the adsorbed BTEX molecules. The monomolecular kinetic equation of photocatalytic process forms (Eq. (1)) [138].

$$r = -\frac{dC}{dt} = k\theta = k\frac{KC}{(1+KC)}$$
(1)

Where r is the reaction rate, C is the concentration of the reactants in the reaction,  $\theta$  is the fraction of the reductant adsorbed to the surface, and k and K are the rate constant and the adsorption equilibrium constant, respectively. Qiu's results had a little bit different from the L-H model owing to the shielding effect which ascribed to the higher concentration of pollutants and the generation of intermediate [135]. Furthermore, the monomolecular kinetic equation would be the firstorder kinetics when the concentration is very small [139]. When it comes to the competitive adsorption of water, the bimolecular kinetic equation is used in the presence of water vapor [137].

$$r = k \frac{K_1 C_1}{1 + K_1 C_1 + K_2 C_2} \tag{2}$$

Where  $K_1$  and  $K_2$  are adsorption equilibrium constants of BTEX and water, and  $C_1$  and  $C_2$  are concentrations of BTEX and water in the gas phase, respectively. Yu's group investigated the effect of adsorption on reaction kinetics during photocatalytic degradation of BTEX by the bimolecular kinetic equation [140]. And the relationship between Langmuir adsorption constant ( $K_1$ ,  $K_2$ ) and  $K_H$  (the Henry's Law constant of VOC) was compared, revealing that there was a positive linear relationship between  $1/K_1$  and  $K_H$ , and a negative relationship between  $1/K_2$  and  $K_H$ . Nguyen *et al.* inferred the apparent rate constants during the whole process according to the L-H model, indicating that the efficiency of VOCs increased at low concentrations [139].

#### 4.2. Adsorption

The adsorption on the catalytic surface is an integral part in photocatalytic technology, which is the beginning of the whole reaction. The adsorption of BTEX on TiO<sub>2</sub> was studied using DFT calculations [141]. There were the aromatic  $\pi$  interactions between BTEX and the surfaces, including BTEX- O3C (tri-coordinated oxygen atoms)-TiO2 and BTEX-Ti<sub>5C</sub> (Penta-coordinated titanium atoms)-TiO<sub>2</sub>. Fig. 6 showed the charge density difference (CDD) between BTEX and TiO2 anatase. By the calculation, the electron density on the  $\pi$  region of aromatic ring close to the surface of all BTEX increased, revealed that there were the aromatic  $\pi$  interaction in the adsorption of BTEX. The state shifted to a smaller energy for the adsorption on the  $\mathrm{Ti}_{5\mathrm{C}}$  in Fig. 6a and the state remained the same for the adsorption on the  $O_{3C}$  indicating that the benzene adsorbent was more stable on  $Ti_{5C}$  than  $O_{3C}$ , which had the same changes as the adsorption of other BTEX. In addition, there was the hybridization between the d state of  $Ti_{5C}$  and carbon p states of benzene. Differently, the methyl hydrogen also played an important influence for xylene on the adsorption process. It was also reported in the literature the adsorption of o-xylene on different surface of TiO<sub>2</sub> suggesting that the stronger adsorption of o-xylene occurred on TiO<sub>2</sub> (001) rather than TiO<sub>2</sub> (101) [35].

Chen' group studied the adsorption process on the SnO<sub>2</sub> surface in dark condition using the *in situ* DRIFTs, as shown in Fig. 7 [142]. The benzene and phenol absorption bands were observed in the adsorption process of benzene (Fig. 7a). As shown in Fig. 6b, there were some peaks certified as toluene, benzyl alcohol, benzaldehyde, and benzoic acid in the toluene adsorption process, inferring that those intermediates were produced by chemical adsorption. In o-xylene adsorption process, it was found that o-xylene, o-diphenol and o-phthaladehyde were detected. The chemical adsorption occurred on the SnO<sub>2</sub> surface forming the preoxidized products. Furthermore, the adsorption energy on the surface of SnO<sub>2</sub> was calculated by DFT (Fig. 7d-f), indicating that the catalyst showed the best adsorption capacity for o-xylene. The results concluded the presence of methyl groups accelerated the adsorption of BTEX. There were benzyl alcohol, benzaldehyde, and benzoic acid in the adsorption process of toluene on Sr<sub>2</sub>Sb<sub>2</sub>O<sub>7</sub> with abundant surface lattice oxygen



**Fig. 6.** The CDD of (a) Benzene- $Ti_{5C}$ -anatase; (b) Benzene- $O_{3C}$ -anatase; (c) Toluene- $Ti_{5C}$ -anatase; (d) Toluene- $O_{3C}$ -anatase; (e) Ethylbenzene- $Ti_{5C}$ -anatase; (f) Ethylbenzene- $O_{3C}$ -anatase; (g) o-Xylene- $Ti_{5C}$ -anatase; (h) o-Xylene- $O_{3C}$ -anatase; (i) m-Xylene- $Ti_{5C}$ -anatase; (j) m-Xylene- $O_{3C}$ -anatase; (k) p-Xylene- $Ti_{5C}$ -anatase and (l) p-Xylene- $O_{3C}$ -anatase. Red is for the negative values (electronic density decreased), and blue is for the positive values (electronic density increased). Isosurfaces of 0.0005 e/bohr<sup>3</sup>.[29].



Fig. 7. The adsorption process of benzene, toluene, and o-xylene in dark condition on SnO<sub>2</sub> surface [142].

[143]. Lin *et al.* [144] researched the adsorption behavior of benzene with the presence of  $H_2O$  on the TiO<sub>2</sub> catalyst. DFT calculations revealed that the adsorption on the TiO<sub>2</sub> surface was strengthened first and then suppressed as  $H_2O$  molecules increased.

# 4.3. Surface reaction

The photocatalytic reactions are complicated and heterogenetic occurring on the surface of catalysts. Generally speaking, the surface

mineralization of  $CO_2$  and  $H_2O$ . In this process, it is inevitable that a lot of intermediates or by-products will be formed [117,145]. It is necessary to study the path and mechanism of photo oxidation reaction of benzene series due to the competitive adsorption of intermediate and by-product [82,106].

reaction of BTEX is a series of redox reactions, eventually leading to the

# 4.3.1. Benzene

A series of the radical trapping experiments demonstrated that the



Fig. 8. (a) Photocatalytic degradation pathway of benzene [48]. (b) Photocatalytic degradation pathway of toluene [81].

superoxide radical ( $\bullet O_2^-$ ), hydroxyl radical ( $\bullet OH$ ) and hole ( $h^+$ ) were the main radicals in the photodegradation of benzene [146]. As shown in Fig. 8a, the intermediates of benzene were observed during the photocatalytic process using the MS [48]. Firstly, benzene interacted with the •OH and  $h^+$  forming phenol (m/z = 93). Benzene was preliminarily attacked by  $h^+$  with the generation of benzene radical cations (•C<sub>6</sub>H<sub>5</sub><sup>+</sup>), and then the  $\bullet C_6H_5^+$  reacted with the surface  $\bullet OH$  to produce phenol [88]. Other possible paths to form the phenol intermediates from the •C<sub>6</sub>H<sub>5</sub><sup>+</sup> have also been reported. The •C<sub>6</sub>H<sub>5</sub><sup>+</sup> was oxide with O<sub>2</sub> or •O<sub>2</sub><sup>-</sup> to produce a peroxide reduced to phenol by  $H^+$  [147]. The phenol was oxidized to benzoquinone (m/z = 93) by  $\bullet O_2^-$  swiftly. Also, there were some peaks identified as p-benzoquinone and o-benzoquinone in some studies, attributing to the further oxidation of phenol [109,148]. Other peaks appeared subsequently identified as malonic acid, oxalic acid, glycolic acid, acetaldehyde, acetic acid, and formaldehyde, which can deduce that the •OH or  $h^+$  opened the benzene ring, forming many organic acids and aldehydes. Finally, all the organic acids and aldehydes were further oxidized to CO<sub>2</sub> and H<sub>2</sub>O. Nevertheless, the highly stable coke deposits might be produced by a free radical polymerization on  $\bullet C_6 H_5^+$ . When the surface  $\bullet OH$  or  $H_2 O$  were insufficient, a number of coke deposits would be formed and block the active sites, resulting in the deactivation of catalyst [88].

Moreover, the aromatic ring might open initially in the photocatalytic process. Under the irradiation with a wavelength less than 200 nm,  $C_6H_6$  can dehydrogenate to generated the unsaturated hydrocarbon [149]. And then, the reduction of such unsaturated hydrocarbon was related to the reactive hydrogen. The presence of nonane, decane, and octane revealed that the generated substances interacted with methyl, leading to the emergence of saturated hydrocarbon identified by the GC–MS [148]. The saturated hydrocarbon was oxidized to organic acids and aldehydes and eventually mineralized to  $CO_2$  and  $H_2O$ .

# 4.3.2. Toluene

In previous researches, many intermediates and by-products were identified by the in situ FTIR spectra and MS in the photocatalytic process of toluene [100,121]. According to the intermediates and byproducts, the possible toluene degradation pathway had been brought up, such as Route A, Route B, Route C (Fig. 8b). In Route A, the toluene was directly oxidized with  $\bullet O_2^-$  forming the benzaldehyde and benzoic acid, which were opened the benzene ring to generate low-molecularweight acids and completely oxidized to CO2 and H2O. In Route B, benzyl was the most important intermediate formed by the reaction of toluene and reactive oxygen species ( $\bullet$ OH or  $\bullet$ O<sub>2</sub>). Then, benzyl possibly was oxidized to benzaldehyde or benzyl alcohol, both of which were eventually oxidized to benzoic acid. Li' group found that the benzoic acid could be ring-opened more easily than the benzaldehyde and benzyl alcohol due to the lowest energy barrier [150]. Furthermore, Dong' group found that the energy barrier for ring-opening of benzoic was lowest, implying that ring opening starting from benzoic acid was observed to be the most favorable for photocatalytic process [151,152]. And the benzyl radicals can react with the toluene forming the dimethyl isophthalate and dimethyl terephthalate [131]. The aromatic compounds were decomposed into CO2 and H2O through a series of oxidation, finally. Jafari et al. [124] and Liu et al. [153] found that the photodegradation path of toluene was consistent with Route B. However, Jafari et al. [124] and Cheng et al. [121] also found the appearance of CO from the GC-MS instrument, attributing to incomplete oxidation of organic matter. In Route C, p-cresol and hydroquinone emerged by the attack of •OH, and the ring-opening reaction took place sequentially resulting in the formation of low-molecular-weight acids, CO<sub>2</sub> and H<sub>2</sub>O. In addition, the photocatalytic process of toluene using the H<sub>3</sub>PW<sub>12</sub>O<sub>40</sub>/ g-C<sub>3</sub>N<sub>4</sub> film-coated optical fibers might have a different path owing to the emergence of p-hydroxybenzoic acid, 2,4-dihydroxybenzoic acid and succinic acid [117].

#### 4.3.3. Ethylbenzene

As reported in some literatures, Ethylbenzene is generally oxidized more easily than benzene and toluene [29,30,154]. The ethyl group on the benzene ring is more reactive than the methyl group [30]. Hinojosa-Reyes's group found the benzaldehyde and acetophenone in the irradiated gas stream generated by direct photo-oxygenation of organic radicals [155]. Similarly, the ozone-assisted photodegradation pathway for ethylbenzene was investigated by GC–MS and IC [156]. There were nine main types of intermediates in ozone-assisted photodegradation of ethylbenzene, including  $\alpha$ -phenylethanol, acetyl benzene,  $\beta$ -phenylethanol, benzyl alcohol, benzaldehyde, benzoic acid, phenol, formic acid and acetic acid. According to the intermediates, the first possible reaction was the oxidation of ethylbenzene by reactive oxygen species forming  $\alpha$ -phenylethanol and  $\beta$ -phenylethanol. The further oxidation was occurred with O<sub>3</sub> to produce acetyl benzene and phenylacetic acid, which can remove a formic acid to obtain the benzyl alcohol. And then, the benzyl alcohol was decarboxylated and opened the aromatic ring obtaining some small organic acids. Finally, they were decomposed into CO<sub>2</sub> and H<sub>2</sub>O [154].

# 4.3.4. Xylene

The methyl number in the benzene ring plays an important role in BTEX decomposition. It is unexpected to discover that the activity of BTEX decomposition and quantum efficiency improved with the increase of methyl number [142]. The xylene has three isomerides, including o-xylene, m-xylene and p-xylene. The possible photocatalytic degradation pathways of the three types of xylenes are roughly similar. Under UV-vis light, the possible path ways of photo-oxidizing xylene are diversiform based on the different intermediates analyzed by GC-MS and IR, and there are usually 4 routes. Route 1, the methyl groups of xylene are attacked by  $\bullet OH$  or  $\bullet O_2^-$  getting 2-benzylalcohol, 2-benzaldehyde, isophthalaldehyde, hydroxybenzoic acid and benzoquinone [35,65,117,154]. And then, those intermediates suffer from ringopening and decomposition completely. Route 2, the xylene is removed a methyl group to form toluene and even further lose the other methyl to form benzene [65,157]. The removed methyl group might react with hydroxyl to produce formaldehyde [87]. Route 3, the xylene could be isomerized into the other xylene and ethylbenzene by the rearrangement [87,157]. Route 4, the xylene can open the ring directly and split into smaller organic molecules, such as acetylene, allylene, 2butane, 2-methyl propene, and so on [87,157]. Those organic molecules are mineralized easily to H<sub>2</sub>O and CO<sub>2</sub>.

# 5. Practical applications and challenges

#### 5.1. Practical applications

The VOCs discharged from multifarious industries have diverse components and complex nature. And exhaust gas generally has a large gas flow rate. The photocatalytic technology may be under restrictions when the degradation occurs in real conditions, such as higher airflow rate, multi-components. Therefore, it is popular for VOCs removal in practical applications to combine other technologies with photocatalytic technology, which exhibits better efficient removal and lower prices.

## 5.1.1. Adsorption-photocatalysis combination

The technology of adsorption-photocatalysis combination is the most common in the volatile organic waste gas treatment (Fig. 9a). The adsorption not only reduces the inlet concentration of photocatalysis, but also decreases the emission of products. Moreover, the technology of adsorption-photocatalysis combination degrades the BTEX to  $CO_2$  and  $H_2O$  which uses the photocatalysts loading on the adsorption materials. On the one hand, the adsorption materials could enrich the concentration of BTEX before the photocatalysis owing to the larger specific surface area and porous structure. On the other hand, the adsorption may lengthen the residence time resulting in more time for photochemical



Fig. 9. (a) The typical process flow diagram of VOCs waste gas treatment in the automobile maintenance industry. (b) A hybrid system combining plasma with photocatalysis for treatment [130].

reactions.

The carbon materials are traditional carriers for the rich porous structure and large surface area, comprised of activated carbon (AC), graphene, carbon nanotubes (CNTs), activated carbon fibers (ACFs), and biochar. The TiO<sub>2</sub> films were coated on AC for toluene degradation resulting that the toluene removal efficiency increased 10 % compared the pure TiO<sub>2</sub> [158]. The interaction effect of AC and TiO<sub>2</sub> loading was studied on toluene removal [159]. The degradation efficiency increased initially and then decreased with the TiO<sub>2</sub> and AC loading increasing attributing to the synergetic adsorption-photodegradation effect. ACFs are ideal support materials with microporous structures widely used in photocatalytic processes [160-162]. The addition of ACFs not only improved the performance of degradation greatly, but also prolonged the service life of photocatalysts. Furthermore, molecular sieve is a kind of porous material with strong adsorption capacity and predominant stability. A series of TiO2 photocatalysts prepared by sol-gel method were germinated in-situ on the surface of ZMS-5 and MCM-41 for benzene degradation [163,164]. The molecular sieve (ZMS-5 and MCM-41) provided high surface area for the dispersion of TiO<sub>2</sub> and accelerated the synergistic adsorption.

# 5.1.2. Plasma-photocatalysis combination

Nowadays, the plasma has attracted extensive attention with superior selectivity and wide range of application. The hard electrons produced by discharging of plasma gas could react with the VOCs in plasma technology. With the assistant of plasma, there are higher concentrations of reactive oxygen species, which could improve the removal efficiency and reduce the emissions of intermediate [165,166]. Benzene was effectively degraded using the non-thermal plasma (NTP) and photocatalytic process in refrigerated food chambers (Fig. 9b) [130]. A synergetic effect between NTP and photocatalysis was observed. In addition, the regeneration of the catalyst by NTP promotes benzene

desorption and the degradation of intermediates. The oxygen plasma was applied for removal of VOCs, accelerating the transfer of electrons and forming more reactive oxygen species [56]. Zhu *et al.*[167] studied the degradation of toluene by F-TiO<sub>2</sub>/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> coupled with dielectric barrier discharge (DBD) NTP. There were abundant high-energy electrons, ions and free radicals in the DBD-NTP, oxidizing and even mineralizing the byproducts to CO<sub>2</sub> and H<sub>2</sub>O.

# 5.1.3. Thermo-photocatalysis combination

The thermo-photocatalysis combination can greatly promote the performance for removing VOCs ascribing the synergistic effect of light energy and heat energy, which is not only conducive to energy saving and improving the stability of the catalyst, but also may improve the selectivity of the target product [127,168]. The photothermal catalytic materials should have a strong whole solar spectral response to achieve the maximum utilization of sunlight. Some researchers have focused mainly on materials responding well to infrared light because the sunlight consists of about 53 % infrared light [169,170]. Combining with the reduced graphene oxide (rGO), the absorption of TiO<sub>2</sub> significantly increased in visible-infrared light, which gave the credit to the electronic interactions between TiO2 and rGO [169]. Yang' group studied the multifunctional catalyst for thermo-catalytic, photocatalytic and photothermo-catalytic oxidation (TCO, PCO and PTO) [90]. The removal efficiency of toluene highly improved in the co-presence of visible-light illumination and heating owing to the lower activation energy and photo-thermal synergy in PTO process. The temperature in toluene removal reduced 40 °C under simulated solar irradiation using CeO<sub>2</sub> modified with Fe, attributing to the synergetic effect in PTO process [171]. The photocatalysis can decrease the accumulation of  $H_2O$  on catalysts, improving the activity and stability of catalysts.

# 5.2. Challenges

The stability and sustainable utilization of catalysts are important indexes for practical application. Some photocatalysts reduce their operational lifetime over long periods of exposure and reuse owing to the dirt of the catalysts surface or the instability of the material itself [130,172]. The photocatalysts prepared by Luu' group exhibited instability for decomposition of p-xylene, which activity dropped tremendously [92]. Importantly, the dirt result from the many recalcitrant intermediates which are more powerfully adsorbed on the catalyst than the reactants. The accumulation of the intermediates not only impedes the light absorption but also occupies the active sites resulting in the lack of reactive oxygen species. And the carbon deposits on the surface gives rise to the deactivation of catalysts gradually. Besides, the dehydroxvlation on the surface might occur in some photocatalysts leading to reducing their activity because of the consumption of hydroxyl radicals [173]. It is necessary to recover the photocatalytic performance by regeneration techniques after the long-term use of photocatalysts. Generally heating and UV cleaning are admittedly effective methods for the regeneration of catalysts. Xu *et al.* [52] found that heating in 20 vol% H<sub>2</sub>/He at 200 °C for 0.5 h could effectively regenerate the catalyst. And different regeneration process under the UV irradiation for 1 h in the clean air was adopted by Wu' group, and the photocatalytic activity can return to the original state [174]. Furthermore, combining the photocatalysis and non-thermal plasma (NTP) can put off the fouling on the catalysts' surface, and the adhibition of NTP without the UV irradiation successfully recovered the performance of photocatalysts [130].

#### 6. Conclusions and perspectives

Photocatalytic technology is considered to be the most promising technologies owing to low-energy, and harmlessness. TiO2, modified with metals, non-metals, or other semiconductors, exhibits superior performance widely noticed and applied to the photocatalytic oxidation of BTEX which not only improves photocatalytic activity, but also prolongs its used life. It is reported that the other metal oxides have a wide band gap and absorption of light, such as ZnO, perovskite oxides, spinel oxides and et al. What's noteworthy is that the hydroxides demonstrated excellent catalytic performance owing to the abundant hydroxyl radicals. In addition, Bi-based photocatalysts, including binary metal oxides, bismuthate and bismuth oxyhalides, and  $C_3N_4$  show the admirable visible light absorption capacity applying to photocatalysis. Furthermore, the operating conditions including the operation mode, initial concentration, airflow rate, temperature, RH and the light were studied adequately. Generally speaking, the photocatalytic process of BTEX mostly follows the L-H model. It is essential to investigate the path of the photocatalytic process for practical application. Although the BTEX have a benzene ring, the degradation pathways differ in details owing to the location and amount of methyl and ethyl groups. In addition, the combination with other technologies and the regeneration of deactivated catalysts are of great significance in practical application.

Although some progresses have been made in the field of photocatalytic degradation about BTEX, there are still many issues that must be solved to meet stringent emission standards in a cost-effective manner. Looking ahead, we consider that the future efforts should be focused on the following aspects.

- (1) The heterostructure photocatalysts show superior application prospects, with significantly reduced band gaps and high electron-hole separation. A promising development about photocatalysts with heterojunction should be considered to make full use of the solar spectrum.
- (2) Fluid dynamics models and other models can be used to simulate the actual working conditions and design a suitable and efficient reactor according to the influence of operating parameters to overcome the issues in the practical application of photocatalytic

degradation of BTEX, such as mass transfer limitations and slow reaction kinetics.

- (3) The theoretical calculation (such as DFT) should be combined with experiments to calculate active sites, adsorption energy, and energy barriers of reactions, which can optimize the catalyst structure to improve the performance, determine the reaction rate control steps to improve the reaction rate, and explore the catalyst deactivation problem to promote its effective regeneration.
- (4) Most experiments were conducted in the laboratory with higher concentrations and simpler compositions, of which the performance and mechanism were probably different from the realistic working conditions. Therefore, more studies should be carried out in term of photodegradation of mixed VOCs with high flow from industrial processes and highly diluted pollutions from indoor environments.
- (5) Developing coupling technologies greatly enhances degradation of BTEX in a more efficient and cost-effective way, including adsorption-photocatalysis, plasma-photocatalysis, and thermophotocatalysis.

#### **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

#### Data availability

No data was used for the research described in the article.

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