

Received: 12 October 2015 Accepted: 08 January 2016 Published: 11 February 2016

OPEN Activity Variation of *Phanerochaete* chrysosporium under Nanosilver **Exposure by Controlling of Different Sulfide Sources**

Zhi Guo^{1,2}, Guiqiu Chen^{1,2}, Lingzhi Liu^{1,2}, Guangming Zeng^{1,2}, Zhenzhen Huang^{1,2}, Anwei Chen³ & Liang Hu^{1,2}

Due to the particular activation and inhibition behavior of silver nanoparticles (AgNPs) on microbes at various concentrations, it's crucial to exploit the special concentration effect in environment. Here, we studied the viability variation of Phanerochaete chrysosporium (P. chrysosporium) under exposure to citrate-coated AgNPs (Citrate-AgNPs) in the presence of different sulfide sources (an inorganic sulfide, NaHS and an organic sulfide, thioacetamide (TAA)). The results indicated that both NaHS and TAA can promote activation of P. chrysosporium by Citrate-AqNPs at a higher concentration, which was initial at toxic level. Treatment with various concentrations of Citrate-AqNPs (0-9 mg/L) demonstrated a maximum activation concentration (MAC) at 3 mg/L. With the increase in sulfide concentration, MAC transferred to higher concentration significantly, indicating the obvious "toxicity to activation" transformation at a higher concentration. Ag+ testing exhibited that variations in sulfide-induced Ag+ concentration $(3-7 \mu q/L Aq^+)$ accounted for the "toxicity to activation" transformation. In addition, the similar results were observed on antibacterial application using Escherichiα coli as the model species. Based on the research results, the application of this transformation in improving antibacterial activity was proposed. Therefore, the antibacterial activity of AgNPs can be controlled, even at concentration, via adjusting for the sulfide concentration.

As a broad-spectrum antimicrobial agent, silver nanoparticles (AgNPs) have been widely used in consumer and medical products. This increased usage of AgNPs translates into increased potential for their release to the environment¹. This situation has attracted considerable interest in toxicity testing^{2,3}, and several studies have exposed AgNPs to various organisms, including bacteria^{4,5}, algae⁶, fungi⁷, Caenorhabditis elegans⁸, zebrafish⁹, and human cells 10,11, to verify their specific manner of toxicity and mechanism. Researchers have demonstrated that AgNPs are preferable candidates for antibiotic drugs; however, their physical and chemical characteristics can be impacted by water chemistry properties such as the presence of some anion ions (S²-, Cl⁻, SO₄²-)^{2,12}.

There are two confirmed patterns of action considered as the main toxicity manner of AgNPs: the release of silver ions from the crystalline core of AgNPs, and the promotion of free radical production 13-16. Although extensive membrane damage has been observed more severely for AgNPs than for ionic $Ag^{+17,18}$, the vast majority of toxicity studies of AgNPs have been conducted with respect to the Ag+ release mechanism. Some other properties such as the AgNP size, surface coating, surface charge, shape, and solubility have also been considered^{2,12,19}. However, the influence of these factors on the toxicity of AgNPs via indirectly affecting silver ion release remains

Despite the great amount attention paid to AgNPs based on toxicity applications^{20–25}, Xiu et al. discovered a new interesting phenomenon that a relatively low concentration (sublethal dose or 12–31% of the minimum lethal concentration for effective Ag+) of AgNPs could stimulate Escherichia coli activity rather than inhibit it⁵. The same result was proposed by our group, in which *Phanerochaete chrysosporium* (*P. chrysosporium*) was

¹College of Environmental Science and Engineering, Hunan University, Changsha 410082, P.R. China. ²Key Laboratory of Environmental Biology and Pollution Control (Hunan University), Ministry of Education, Changsha 410082, P.R. China. ³College of Resources and Environment, Hunan Agricultural University, Changsha 410128, P.R. China. Correspondence and requests for materials should be addressed to G.C. (email: gqchen@hnu.edu.cn) or G.Z. (email: zgming@hnu.edu.cn)

target organisms	reported maximum activation point (mg/L)	optimal activation range ^a (mg/L)	AgNPs category	reference
E. coli	2.2	1.0-4.0	PEG-AgNPs-3 nm	5
	1.8	0-3.5	PEG-AgNPs-5 nm	
	2	0-4.5	PEG-AgNPs-11 nm	
E. coli	16.4	5.0-35.0	PVP-AgNPs-20 nm	
	5.7	0-10.0	PVP-AgNPs-40 nm	5
	6.7	10.0-30.0	PVP-AgNPs-80 nm	
P. chrysosporium	1	0.1-5.0	Citrate-AgNPs-24nm	26

Table 1. Proposed AgNPs hormesis effect to microbe. ^aestimated from the figure of the literature.

activated by a low concentration (the range is about $0.1-5.0\,\text{mg/L}$) of AgNPs, which promoted its removal capacity to cadmium ions (Table 1)²⁶. This result is a great threat to potential applications of AgNPs to exploit the predominant antibacterial action of its products. These findings suggest that in some cases, the antifungal products may stimulate microbial growth but not inhibition as expected. Hence, it is necessary to verify the specific stimulatory effect of AgNPs under both laboratory and real conditions.

Sulfides are widely present in the environment and biological systems, and display strong chemical activity to other matter $^{27-31}$. Previous studies have indicated that the sulfidation of Ag surfaces is likely to occur when in contact with various S-bearing molecules². Sulfidation can impact the thermal and electrical conductivity of silver, and even result in the formation of a new shell (such as Ag₂S) on the surface, which strongly affects the toxicity of AgNPs^{32,33}. Bone *et al.* demonstrated that more Ag was present as Ag₂S in the absence of plants³⁴. Liu *et al.* demonstrated that sulfidation of AgNPs occurs via two different mechanisms depending on the sulfide concentration³⁵. At high sulfide concentrations, sulfidation occurs by direct conversion of AgNPs to Ag₂S-NPs through a solid—fluid reaction, whereas at lower sulfide concentrations, oxidative dissolution and precipitations seems to be prevalent. Hence, it is necessary to evaluate the effect of sulfides on the toxicity performance of AgNPs in consideration of its antifungal application.

P. chrysosporium, as the representative species of white-rot fungi, has been extensively used for its ability to degrade a wide range of organic substrates and to absorb heavy metals in wastewater treatment field^{36–38}. The widely use of AgNPs inevitably results in its access to biological water treatment system. Hence, exploitation of the toxic effect of AgNPs on *P. chrysosporium* is an important component in the process of wastewater treatment. In this article, we demonstrated that environmental sulfide induced "toxicity to activation" transformation of citrate-coated AgNPs (Citrate-AgNPs). In the absence of sulfide, a maximum activation concentration (MAC) of Citrate-AgNPs at 3 mg/L was observed. Sulfide addition promoted the initial toxicity of Citrate-AgNPs transformation to activation. Increase of sulfide concentration resulted in higher MAC transfer. Based on this result, we propose a method for controlling the antibacterial activity of Citrate-AgNPs, even at low concentration (3 mg/L), via adjusting for sulfide concentration effectively.

Results and Discussion

Synthesis and Characterization of Citrate-AgNPs. Citrate-AgNPs were synthesized according to the method described by Liu *et al.* with slight modifications^{39,40}. Figure S1 shows that Citrate-AgNPs are spherical in shape with a mean size of 15.09 ± 2.21 . The Citrate-AgNPs were found to be nonaggregating in both deionized water and the minimal medium used for assays. The Z-potential was demonstrated to be -47.4 ± 0.7 mV and a typical plasma resonance absorption peak at 396 nm of AgNPs was observed (Fig. S2).

MAC Transfer and "Toxicity to Activation" Transformation can be Triggered by Sulfide. To the best of our knowledge, organic and inorganic matters exhibit different biocompatibilities to cells, which may translate into different effects on the toxicity performance of AgNPs⁴¹. In this study, the inorganic sulfide source NaHS and the organic sulfide source thioacetamide (TAA) were used as the sulfide sources. To clarify the effect of sulfides on the stimulatory action of Citrate-AgNPs to microbes, we defined the maximum activation concentration of the tested Citrate-AgNPs as the MAC. The MAC of Citrate-AgNPs to P. chrysosporium was found to be 3 mg/L without the addition of sulfide (Fig. 1). This result supports previous findings of our group, in which the MAC of Citrate-AgNPs to P. chrysosporium was determined to be in the range of $1-5 \,\text{mg/L}^{26}$. As shown in Fig. 1, the addition of aqueous NaHS promoted substantial MAC changes. The MAC transferred to a higher concentration with the increase of NaHS concentration from 0 to $200\,\mu M$ as the sulfide source. The transfer degrees induced by various concentration of sulfides (TAA and NaHS) (0, 25, 50, 100, and 200 μM) were 0, 1, 2, 4, and 5 mg/L, respectively. Interestingly, we found that the maximum viability of *P. chrysosporium* at the MAC appeared at 100 μM NaHS but not at 200 μM NaHS; thus, the maximum viability of *P. chrysosporium* at the MAC was not proportional to the NaHS concentration. This different relationship between MAC transfer and maximum viability of *P. chrysosporium* at MAC indicates that the change in MAC transfer was not directly caused by an effect of NaHS on promoting *P. chrysosporium* survival.

Figure 2 shows that TAA produced similar effects to NaHS in terms of the MAC transfer. However, the maximum viability of *P. chrysosporium* at MAC was different, with the best survival observed at a concentration of $50\,\mu\text{M}$ TAA. This result further demonstrates the irrelevance of MAC transfer to changes in *P. chrysosporium* survival promoted by sulfides. In addition, NaHS and TAA showed similar effects on the toxicity

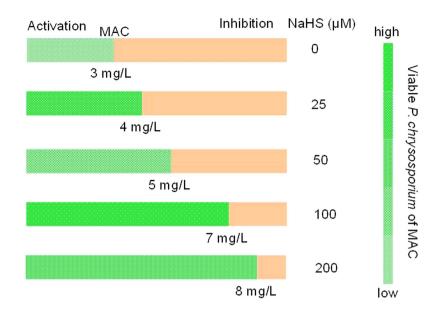


Figure 1. MAC transfer and viable P. chrysosporium of MAC induced by NaHS as the sulfide source.

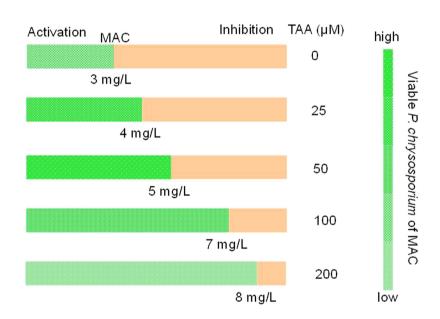
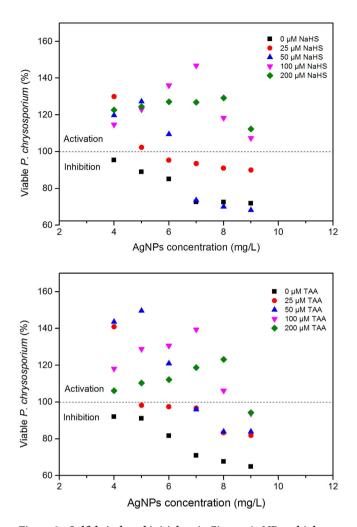


Figure 2. MAC transfer and viable P. chrysosporium of MAC induced by TAA as the sulfide source.

of the Citrate-AgNPs. Based on these results, the maximum MAC of $100\,\mu\text{M}$ NaHS, $7\,\text{mg/L}$ Citrate-AgNPs and $50\,\mu\text{M}$ TAA, $5\,\text{mg/L}$ Citrate-AgNPs were used in subsequent assays.

The results described above demonstrated that the sulfides source (both inorganic and organic) induced the activation point transfer of Citrate-AgNPs to high concentration (from 3 mg/L to 8 mg/L with the sulfide concentration added to $200\,\mu\text{M}$). However, high concentration (>4 mg/L for *P. chrysosporium*) consistently showed excellent inhibition to cell viability in the absence of sulfide. As shown in Fig. 3, this trend resulted in initial toxic Citrate-AgNPs at higher concentration to transfer to an activation effect with the increase of sulfide. This interesting "toxicity to activation" transformation phenomenon prompted us to further evaluate the effect of AgNPs on bacteriostasis.

Explanation of the Transfer Phenomenon. To discern the contribution of sulfide to the toxicity of Citrate-AgNPs, NaHS, TAA, and equivalent hydrolysis products of TAA, including CH₃COO⁻ and NH₃, were tested with and without Citrate-AgNPs in terms of their effects on the transfer degree of MAC and *P. chrysos-porium* activation ratio (Fig. 4). The results showed that CH₃COO⁻ and NH₃ had no effect on the MAC transfer to higher concentration (CH₃COO⁻ and NH₃ cause MAC transfer to lower concentration at 1 mg/L and 2 mg/L, respectively), indicating the essential role of sulfide in promoting the transfer of Citrate-AgNPs. With respect



 $Figure \ 3. \ Sulfide \ induced \ initial \ toxic \ Citrate-AgNPs \ at \ higher \ concentration \ transfer \ to \ activation.$

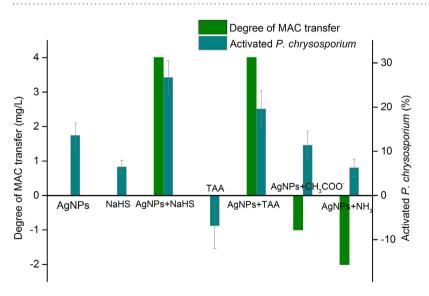


Figure 4. Degree of transfer and P. chrysosporium activation at MAC induced by relevant factors . The concentration was 7 mg/L for AgNPs in the P. chrysosporium viability tests, and 100 μ M for NaHS, TAA, CH₃COO⁻ and NH₃ for all the needed tests.

to activation, Citrate-AgNPs and NaHS simultaneously exhibited stimulation effects to *P. chrysosporium*, and CH₃COO⁻ and NH₃ had some effect on the toxicity of Citrate-AgNPs. These results suggest that sulfide itself has

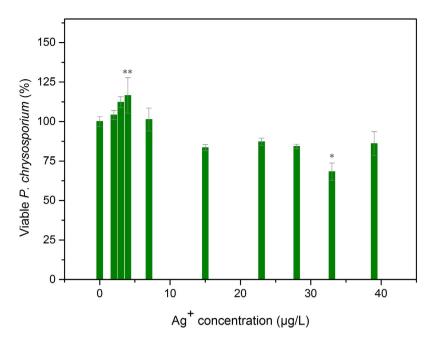


Figure 5. Survival of resting P. chrysosporium cells in buffer solution after 12 h exposure to AgNO3. One asterisk represents significant decrease in viability (p < 0.05) relative to unexposed control. A significant stimulatory effect suggestive of hormesis was observed of all treatments, as indicated by two asterisks.

no effect on the activation of *P. chrysosporium* by Citrate-AgNPs. This may be attributed to the indirect action of sulfide on the toxicity of Citrate-AgNPs. Because the release of Ag⁺ from AgNPs is thought to be one of the main mechanisms governing the toxicity of AgNPs, the MAC would be expected to transfer to a high concentration in the presence of sulfide because of the formation of relatively insoluble Ag₂S. In the solution, AgNPs easily release Ag⁺ in an aerobic environment (equilibrium constant $K = 1 \times 10^{16}$, calculated from the ΔG^0 value of -91.3 kJ/mol at 298 K):³⁹

$$2Ag_{(s)} \ + \ \frac{1}{2}O_{2(aq)} + \ 2H^{+}_{(aq)} \ \rightleftharpoons \ 2Ag^{+}_{(aq)} + 2H_{2}O_{(l)} \eqno(1)$$

Sulfides react with Ag⁺ of AgNPs surface to form of Ag₂S as demonstrated by Levard *et al.*⁴², Normally, Ag₂S dissolves Ag⁺ via the following process. The solubility product constant $K_{\text{sp-Ag2S}}$ is 5.92×10^{-51} for the dissolution process⁴³, indicating the lower solubility and toxicity of Ag₂S^{42,44-47}.

$$Ag_2S + 2H^+ \rightleftharpoons 2Ag^+ + H_2S_{(aq)} \tag{2}$$

The lower solubility product constant of Ag₂S indicates that sulfidation could effectively lower the amount of Ag⁺ released and thus reduce the toxicity of Ag₂S compared to that of AgNPs, accounting for the MAC transfer.

The survival of resting *P. chrysosporium* cells in 2 mM NaHCO₃ buffer solution was activated in the presence of $3-7 \mu g/L$ Ag + after 12 h exposure (the lowest stimulation was less than $3 \mu g/L$ in some cases) (Fig. 5). Ag⁺ concentration ~4 $\mu g/L$ exhibited a significant stimulatory effect compared to the unexposed control group.

The Ag^+ residue in the solution was determined to be conformed to the activation concentration $(3-7\mu g/L)$ for Ag^+ after exposure to 7 mg/L Citrate-AgNPs, $100\,\mu$ M NaHS (Fig. 6), which was the condition for the highest viability of *P. chrysosporium* at MAC, as shown in Fig. 1. Simultaneously, the addition of 5 mg/L Citrate-AgNPs and $50\,\mu$ M TAA showed an optimal stimulation concentration of Ag^+ ($3\,\mu g/L$) (Fig. 7). Higher and lower sulfide concentrations would cause the Ag^+ residue to be outside of the activation range. These Ag^+ concentrations range well explained the variation in *P. chrysosporium* viability at MAC and the effect of sulfide on Ag^+ release. This finding is in line with the interpretation that the toxicity of AgNPs is mainly caused by Ag^+ release. A significant decrease in the surface plasma resonance of Citrate-AgNPs was also observed after the addition of sulfide source (Figs S3 and S4); that may be attributed to stable material production on the surface (Ag_2S , as proposed above) Ag^+ 0. As a result, the Ag^+ 1 release was limited (Figs 6 and 7), the toxicity process of AgNPs based on Ag^+ 1 was interfered. These results together demonstrate that the sulfide did indeed have an effect on Ag^+ 1 release.

Tests of Antibacterial Activity of Citrate-AgNPs and Antifungal/Antibacterial Activity Biologically Synthesized AgNPs. To assess the effect of sulfide on Citrate-AgNPs toxicity on bacteria, gram-negative bacterium *Escherichia coli* (*E. coli*) was employed for the toxicity assays. As shown in Fig. S5, sulfide induced the analogous trend in antibacterial activity. These results indicate that the methods could be used in antibacterial application simultaneously.

In addition, we tested the effect of sulfide on biologically synthesized AgNPs toxicity on *P. chrysosporium* and *E. coli* (Table S1). No obvious MAC transfer was observed on this biologically synthesized AgNPs. The S–H

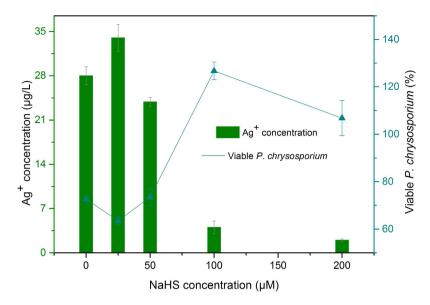


Figure 6. The concentration of residue Ag+ and viable P. chrysosporium at MAC with various NaHS. The concentration of Citrate-AgNPs was 7 mg/L.

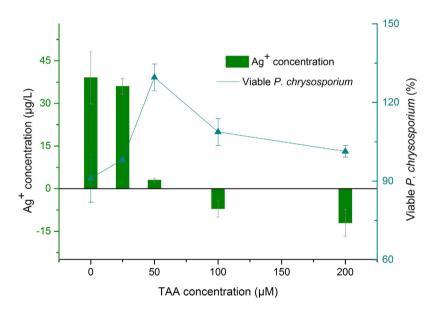
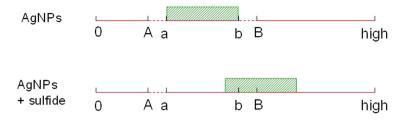


Figure 7. The concentration of residue Ag+ and viable P. chrysosporium at MAC with various TAA. The concentration of Citrate-AgNPs was 5 mg/L.

groups of the AgNPs surface forming an Ag–S bond as demonstrated by Sanghi *et al.*⁴⁸, has a similar effect of TAA and NaHS, which may cause the unobvious MAC transfer.

Potential Application of MAC Transfer in Antibacterial Use. A low dose of AgNPs has a hormesis effect on microbial and animal cell viability (Table S2). This effect has been interpreted as overcompensation for the exposure, because low doses of toxicants can activate the repair mechanisms of cells against the toxicant^{5,49}. As shown in Table S2, the hormesis effect normally occurred in the low concentration range of AgNPs, which was regarded as an effective district for the hormesis effect. Lower and higher concentrations resulted in inhibition of cell viability.

To illustrate the hormesis effect more clearly, we constructed the graph presented in Fig. 8. Under certain conditions, the concentration region [a, b] was considered to be the effective district of the hormesis effect, whereas [0, A] and [B, high] were regarded as the inhibition districts. According to the literature, [A, a] and [b, B] likely represent uncertain districts for unreported activation or even inhibition effects^{50,51}. Figure 8 indicates that the region [a, b] is disadvantageous to the antibacterial effect of AgNPs. External action would be needed to maintain the concentration within the region of [0, A] or [B, high] to ensure the effective antibacterial activity of AgNPs.



[0, A] and [B, High]: Inhibition region [a, b]: Acti∨ation region

Figure 8. chematic for hormesis effect theory of AgNPs. The region of [0, high] indicate the concentration of AgNPs from 0 to higher. The slash box transfer to higher concentration, which indicates the activation region transferring to higher concentration after addition of sulfide (cysteine or Na₂S).

As described above, sulfide can induce the MAC transfer to a higher concentration, shifting the initial low activation or even inhibition concentration ([B, high]) toward high stimulation (Fig. S6). In other words, the main toxic and activation effects of AgNPs are attributed to the released Ag^+ concentration. Hence, the addition of sulfide could lead to a change to inhibition when in its activated state ([a, b]) due to the certain range of the activation district. The usual concentration of AgNPs used in antibacterial agents is low ($<5.0\,\text{mg/L}$)⁵², so that the antibacterial activity is located at the stimulation district. Therefore, sulfide addition could effectively control the antibacterial ability of AgNPs. For example, when point A was 2.5 mg/L and point B was 3.5 mg/L for AgNPs concentration, the initial MAC was 3 mg/L AgNPs. The simple addition of $25\,\mu$ M TAA would cause a MAC transfer to 4 mg/L (Fig. 3). Under the precondition of a uniform sulfidation rate, point A and B would change to 3.5 mg/L and 4.5 mg/L, respectively. The initial MAC of 3 mg/L is therefore transferred to inhibition from the activation state. With continuous addition of sulfide, point A would further transfer to a higher concentration, retaining a larger inhibition range of [0, A]. This proposed strategy could be effective for bacteriostasis applications to enhance the sterilizing effect of AgNPs at a concentration of antimicrobial reagents within stimulation region.

Although this study was conducted with the addition of sulfide, our results may provide an effective approach for effectively controlling and utilizing the hormesis effect of AgNPs in antibacterial applications. Some substances may also interfere with AgNPs bacteriostasis. To avoid activation in the sterilization process, a new convenient method may be developed based on this result.

Conclusions

In summary, this study proposed the MAC transfer of Citrate-AgNPs under the induction of inorganic sulfide source (NaHS) or organic sulfide source (thioacetamide, TAA). Sulfide (TAA or NaHS) addition promoted the MAC transfer to higher AgNPs concentration, which results in the initial inhibition concentration of Citrate-AgNPs changing to activation as "toxicity to activation" transformation pattern. Higher sulfide concentration results in larger transfer of the site. The best activation concentration of sulfide was $100\,\mu\text{M}$ for NaHS and $50\,\mu\text{M}$ for TAA, which achieved maximal viability of *P. chrysosporium* at ~126.7% and ~129.6%, respectively. In addition, the application of this BSP transfer effect in improving antibacterial activity was proposed. The antibacterial activity of AgNPs can be controlled, even at low concentration, via adjusting for the MAC transfer effectively.

Methods

Synthesis of Citrate-AgNPs. A 59.8 mL solution containing 0.6 mM trisodium citrate and 0.4 mM NaBH₄ was prepared in double distilled water and stirred vigorously in an ice bath. Upon addition of 0.48 mL, 23.5 mM AgNO₃ under stirring, the colour of solution turned yellow, indicating the formation of Citrate-AgNPs. After 3 h of additional stirring at room temperature, the soluble byproducts were removed by centrifugal ultrafiltration (molecular weight cutoff of 1000), and the Citrate-AgNPs were washed with double distilled water. The morphology and EDAX spectrum are provided in Fig. S7.

Phanerochaete chrysosporium identification and cultivation. To confirm *P. chrysosporium* isolation, the genomic DNA of *P. chrysosporium* was extracted using the Ezup Column Fungi Genomic DNA Purification Kit following the manufacturer's instructions (Sangon Biotech Co., Ltd., Shanghai, China). The 18S rDNA gene sequence of the genomic DNA was amplified by PCR using primer sets of NS1 (5′-GTAGTCATATGCTTGTCTC-3′) and NS6 (5′-GCATCACAGACCTGTTATTGCCTC-3′). PCR amplification was performed in a 25 μL reaction mixture containing 0.5 μL of each primer (10 μM), 0.5 μL of template DNA (20–50 ng/μL), 0.2 μl of Ex Taq (5 Uμl⁻¹), 2.5 μL of 10 × PCR buffer with mg²⁺, 1.0 μL of dNTP (2.5 mM each), and sterile distilled water to a final volume of 25 μL. PCR cycling was performed under the following conditions: an initial denaturing step at 94 °C for 4 min, followed by 30 cycles of denaturation at 94 °C for 45 s, annealing at 55 °C for 45 s, and elongation at 72 °C for 1 min. Finally, an extension step was performed at 72 °C for 10 min. PCR products were verified by agarose gel electrophoresis (1.0% weight/volume agarose) with ethidium bromide staining and visualized using an ultraviolet (UV) transilluminator. After that, the target gene fragments were purified using SanPrep Column DNA Gel Extraction Kit (Sangon Biotech Co., Ltd.) in accordance with the manufacturer's instructions. And then, the purified products were sequenced by the Sangon Biotech Co., Ltd. The 18S

rDNA gene sequences obtained were analyzed with the BLAST program of the GenBank database at the National Center for Biotechnology Information (NCBI) website (http://www.ncbi.nlm.nih.gov/). The results show that the 18S rDNA sequences of *P. chrysosporium* shared 100% and 99% similarity with *Phanerochaete* sp. Y6 (accession No. DQ438911) and *P. chrysosporium* KSR2 (accession No. KJ606692) respectively.

P. chrysosporium strain BKMF-1767 (CCTCC AF96007) used in this study was obtained from the China Center for Type Culture Collection (Wuhan, China). Stock cultures were maintained on malt extract agar slants at $4\,^{\circ}$ C. Aqueous suspensions of fungal spores were inoculated into Kirk's liquid culture medium and incubated at $37\,^{\circ}$ C in an incubator according to previous report⁵³. After 2 days of growth in liquid culture, the mycelia were harvested by centrifugation at $10000\times g$ for 5 min, washed three times with 2 mM sodium bicarbonate buffer solution, and resuspended in the same buffer to make the *P. chrysosporium* stock solution. Sodium bicarbonate can buffer the system at relatively low ionic strength (which promotes nanoparticle coagulation) compared to other bacteria media, and was chosen as the exposure medium to avoid ligands that could bind with $Ag^+/AgNPs$ and promote precipitation or other confounding effects ($AgNPs/Ag^+$) toxicity assays in this work were below Ag_2CO_3 precipitation potential ($Ksp=0.81\times10^{-12})^{5,13}$.

Dose-response assay of Citrate-AgNPs on *P. chrysosporium*. Equivalent mycelia $(0.4\,\mathrm{g})$ were added respectively into test tubes to achieve an identical cell concentration. Before toxicity tests, Citrate-AgNPs stock solution was diluted in $2\,\mathrm{mM}$ sodium bicarbonate buffer to obtain different concentrations. Aqueous Citrate-AgNPs and sulfide were mixed and equilibrate with *P. chrysosporium* in order to begin the toxicity test. After $12\,\mathrm{h}$, mycelia were centrifugation to remove residual AgNPs and Ag⁺.

Cell viability assays were operated according to Luo *et al.* and Chen *et al.* 54,55 , 0.2 g *P. chrysosporium* pellets were mixed with 1 mL MTT solution (5 g/l) and incubated at 50 °C for 1 h. The reaction was stopped by adding 0.5 mL hydrochloric acid (1 M) to the mixture. The mixture was centrifuged (10,000 \times g, 5 min), the supernatant was discarded, and the pellets were agitated in 6 ml propan-2-ol for 2 h at 25 °C. The centrifugation process was repeated and the absorbance of the supernatant was recorded at 534 nm using spectrophotometer (Model UV-2550, Shimadzu, Japan).

For Ag^+ testes, the solution was ultrafiltration centrifuged at $10000 \times g$. The Ag^+ concentration of filtrate was then determined using ICP-OES (IRIS Intrepid II XSP, Thermo Electron Corporation, USA)

Statistical analyses. Whether differences between treatments were statistically significant was determined using Student's t test at the 95% confidence level (Statistical Program for Social Sciences 19.0). All measurements are reported as mean \pm one standard deviation with three replicates.

Associated Content. SUPPORTING INFORMATION AVAILABLE: Methods of dose-response assay of Citrate-AgNPs on *E. coli*, biosynthesis of AgNPs, and dose-response assay of biologically synthesized AgNPs on *P. chrysosporium* and *E. coli*. TEM, UV-vis characterization of Citrate-AgNPs (Figs S1, S2). Sulfide induced decrease in the surface plasma resonance of Citrate-AgNPs (Figs S3, S4). Sulfide effect of Citrate-AgNPs toxicity on *E. coli* (Fig. S5) and biologically synthesized AgNPs on *P. chrysosporium* and *E. coli* (Fig. S6). Graph for sulfide effect on AgNPs toxicity of initial inhibition state (Fig. S6). HAADF-STEM image of Citrate-AgNPs and EDAX spectrum (Fig. S7). Development of AgNPs hormesis effect (Table S2).

References

- Chambers, B. A. et al. Effects of Chloride and Ionic Strength on Physical Morphology, Dissolution, and Bacterial Toxicity of Silver Nanoparticles. Environ. Sci. Technol. 48, 761–769 (2014).
- Levard, C., Hotze, E. M., Lowry, G.V. & Brown, Jr., G. E. Environmental Transformations of Silver Nanoparticles: Impact on Stability and Toxicity. Environ. Sci. Technol. 46, 6900 – 6914 (2012).
- 3. Eckhardt, S., Brunetto, P. S., Gagnon, J., Priebe, M. & Giese, B. Nanobio Silver: Its Interactions with Peptides and Bacteria, and Its Uses in Medicine. *Chem. Rev.* 113, 4708–4754 (2013).
- 4. Loo, S. L. et al. Bactericidal Mechanisms Revealed for Rapid Water Disinfection by Superabsorbent Cryogels Decorated with Silver Nanoparticles. Environ. Sci. Technol. 49, 2310–2318 (2015).
- Xiu, Z. M., Zhang, Q. B., Puppala, H. L., Colvin, V. L. & Alvarez, P. J. J. Negligible Particle-Specific Antibacterial Activity of Silver Nanoparticles. Nano Lett. 12, 4271–4275 (2012).
- Scanlan, L. D. et al. Silver Nanowire Exposure Results in Internalization and Toxicity to Daphnia magna. ACS Nano 7, 10681–10694
 (2013).
- 7. Navarro, E. *et al.* Environmental Behavior and Ecotoxicity of Engineered Nanoparticles to Algae, Plants, and Fungi. *Ecotoxicology* 17, 372–386 (2008).
- 8. Yang, X. et al. Mechanism of Silver Nanoparticle Toxicity is Dependent on Dissolved Silver and Surface Coating in Caenorhabditis elegans. Environ. Sci. Technol. 46, 1119–1127 (2012).
- Asharani, P. V., Wu, Y. L., Gong, Z. & Valiyaveettil, S. Toxicity of Silver Nanoparticles in Zebrafish Models. Nanotechnology 19, 255102 (2008).
- Lu, W. et al. Effect of Surface Coating on the Toxicity of Silver Nanomaterials on Human Skin Keratinocytes. Chem. Phys. Lett. 487, 92–96 (2010).
- 11. Asharani, P. V., Mun, G. L. K., Hande, M. P. & Valiyaveettil, S. Cytotoxicity and Genotoxicity of Silver Nanoparticles in Human Cells. ACS Nano 3, 279—290 (2009).
- 12. Tejamaya, M., Römer, I., Merrifield, R. C. & Lead, J. R. Stability of Citrate, PVP, and PEG Coated Silver Nanoparticles in Ecotoxicology Media. *Environ. Sci. Technol.* 46, 7011–7017 (2012).
- Xiu, Z. M., Ma, J. & Alvarez, P. J. J. Differential Effect of Common Ligands and Molecular Oxygen on Antimicrobial Activity of Silver Nanoparticles versus Silver Ions. Environ. Sci. Technol. 45, 9003–9008 (2011).
- Ivask, A. et al. Toxicity Mechanisms in Escherichia coli Vary for Silver Nanoparticles and Differ from Ionic Silver. ACS Nano 8, 374–386 (2014).
- 15. Unrine, J. M., Colman, B. P., Bone, A. J., Gondikas, A. P. & Matson, C. W. Biotic and Abiotic Interactions in Aquatic Microcosms Determine Fate and Toxicity of Ag Nanoparticles. Part 1. Aggregation and Dissolution. *Environ. Sci. Technol.* 46, 6915–6924 (2012).
- Wang, H. et al. Toxicity, Bioaccumulation, and Biotransformation of Silver Nanoparticles in Marine Organisms. Environ. Sci. Technol. 48, 13711–13717 (2014).

- Hwang, E. T. et al. Analysis of the Toxic Mode of Action of Silver Nanoparticles Using Stress-specific Bioluminescent Bacteria. Small 4, 746–750 (2008).
- 18. Ruparelia, J. P., Chatterjee, A. K., Duttagupta, S. P. & Mukherji, S. Strain Specificity in Antimicrobial Activity of Silver and Copper Nanoparticles. *Acta Biomater.* 4, 707–716 (2008).
- 19. Ma, R. et al. Size-controlled Dissolution of Organic-coated Silver Nanoparticles. Environ. Sci. Technol. 46, 752-759 (2012).
- Ivask, A. et al. DNA Melting and Genotoxicity Induced by Silver Nanoparticles and Grapheme. Chem. Res. Toxicol. 28, 1023–1035 (2015).
- Wang, L. et al. Use of Synchrotron Radiation-analytical Techniques to Reveal Chemical Origin of Silver-nanoparticle Cytotoxicity. ACS Nano 9, 6532–6547 (2015).
- 22. Chen, L. Q. et al. Nanotoxicity of Silver Nanoparticles to Red Blood Cells: Size Dependent Adsorption, Uptake, and Hemolytic Activity. Chem. Res. Toxicol. 28, 501–509 (2015).
- Srivastava, V., Gusain, D. & Sharma, Y. C. Critical Review on the Toxicity of Some Widely Used Engineered Nanoparticles. Ind. Eng. Chem. Res. 54, 6209–6233 (2015).
- 24. Zhao, X. & Ibuki, Y. Evaluating the Toxicity of Silver Nanoparticles by Detecting Phosphorylation of Histone H3 in Combination with Flow Cytometry Side-scattered Light. *Environ. Sci. Technol.* **49**, 5003–5012 (2015).
- Khan, F. R. et al. Accumulation Dynamics and Acute Toxicity of Silver Nanoparticles to Daphnia magna and Lumbriculus variegatus: Implications for Metal Modeling Approaches. Environ. Sci. Technol. 49, 4389–4397 (2015).
- Zuo, Y. et al. Transport, Fate, and Stimulating Impact of Silver Nanoparticles on the Removal of Cd(II) by Phanerochaete chrysosporium in Aqueous Solutions. J. Hazard. Mater. 285, 236–244 (2015).
- Tuček, J., Kemp, K. Č., Kim, K. S. & Zbořil, R. Iron-oxide-supported Nanocarbon in Lithium-ion Batteries, Medical, Catalytic, and Environmental Applications. ACS Nano 8, 7571–7612 (2014).
- 28. Ribeiro, D. C., Martins, G., Nogueira, R. & Brito, A. G. Mineral Cycling and pH Gradient Related with Biological Activity under Transient Anoxic–Oxic Conditions: Effect on P Mobility in Volcanic Lake Sediments. *Environ. Sci. Technol.* 48, 9205–9210 (2014).
- Kong, H. et al. Hydrogen Sulfide Detection Based on Reflection: from A Poison Test Approach of Ancient China to Single-Cell Accurate Localization. Anal. Chem. 86, 7734

 – 7739 (2014).
- 30. Yan, Y. et al. Molecularly Engineered Quantum Dots for Visualization of Hydrogen Sulfide. ACS Appl. Mater. Interfaces 7, 3547–3553 (2015)
- 31. Guo, K. et al. Optimal Packing of a Rotating Packed Bed for H,S Removal. Environ. Sci. Technol. 48, 6844-6849 (2014).
- 32. Graedel, T. E., Franey, J. P., Gualtieri, G. J., Kammlott, G. W. & Malm, D. L. On the Mechanism of Silver and Copper Sulfidation by Atmospheric H₂S and OCS. *Corros. Sci.* 25, 1163–1180 (1985).
- 33. Franey, J. P., Kammlott, G. W. & Graedel, T. E. The Corrosion of Silver by Atmospheric Sulfurous Gases. *Corros. Sci.* 25, 133–143 (1985).
- 34. Bone, A. J. et al. Biotic and Abiotic Interactions in Aquatic Microcosms Determine Fate and Toxicity of Ag Nanoparticles: Part 2—Toxicity and Ag Speciation. Environ. Sci. Technol. 46, 6925—6933 (2012).
- 35. Liu, J., Pennell, K. G. & Hurt, R. H. Kinetics and Mechanisms of Nanosilver Oxysulfidation. *Environ. Sci. Technol.* **45**, 7345–7353 (2011)
- 36. Zeng, G. M. et al. Responses of Phanerochaete chrysosporium to Toxic Pollutants: Physiological Flux, Oxidative Stress, and Detoxification. Environ. Sci. Technol. 46, 7818–7825 (2012).
- 37. Huang, D. L. et al. Mycelial Growth and Solid-state Fermentation of Lignocellulosic Waste by White-rot Fungus Phanerochaete chrysosporium under Lead Stress. Chemosphere 81, 1091–1097 (2010).
- 38. Yu, M. et al. Influence of Phanerochaete chrysosporium on Microbial Communities and Lignocellulose Degradation during Solid-state Fermentation of Rice Straw. Process Biochem. 44, 17–22 (2009).
- 39. Liu, J. & Hurt, R. H. Ion Release Kinetics and Particle Persistence in Aqueous Nano-silver Colloids. *Environ. Sci. Technol.* 44, 2169–2175 (2010).
- 40. Guo, Z. et al. Ultrasensitive Detection and Co-stability of Mercury(II) Ions Based on Amalgam Formation with Tween 20-stabilized Silver Nanoparticles. RSC Adv. 4, 59275–59283 (2014).
- 41. González, V., Díez-Ortiz, M., Simón, M. & van Gestel, C. A. M. Assessing the Impact of Organic and Inorganic Amendments on the Toxicity and Bioavailability of a Metal-contaminated Soil to the Earthworm *Eisenia andrei. Environ. Sci. Pollut. Res.* 20, 8162–8171 (2013).
- 42. Levard, C. et al. Sulfidation Processes of PVP-coated Silver Nanoparticles in Aqueous Solution: Impact on Dissolution Rate. Environ. Sci. Technol. 45, 5260–5266 (2011).
- 43. Lide, D. R. Handbook of Chemistry and Physics; CRC Press: Boca Raton (2009).
- 44. Suresh, A. K. et al. Monodispersed Biocompatible Silver Sulfide Nanoparticles: Facile Extracellular Biosynthesis using the γ-proteobacterium, Shewanella oneidensis. Acta Biomater. 7, 4253–4258 (2011).
- 45. Nowack, B. Nanosilver revisited downstream. Science 330, 1054–1055 (2010).
- 46. Fabrega, J., Luoma, S. N., Tyler, C. R., Galloway, T. S. & Lead, J. R. Silver Nanoparticles: Behaviour and Effects in the Aquatic Environment. *Environ. Int.* 37, 517–531 (2011).
- 47. Choi, O. *et al.* Role of Sulfide and Ligand Strength in Controlling Nanosilver Toxicity. *Water Res.* **43**, 1879–1886 (2009).
- Sanghi, R. & Verma, P. Biomimetic Synthesis and Characterisation of Protein Capped Silver Nanoparticles. Bioresour. Technol. 100, 501–504 (2009).
- 49. Calabrese, E. J. Overcompensation Stimulation: A Mechanism for Hormetic Effects. Crit. Rev. Toxicol. 31, 425-470 (2001).
- 50. Arora, S., Jain, J., Rajwade, J. M. & Paknikar, K. M. Cellular Responses Induced by Silver Nanoparticles: *In Vitro* Studies. *Toxicol. Lett.* **179,** 93–100 (2008).
- 51. Iavicoli, I., Calabrese, E. J. & Nascarella, M. A. Exposure to Nanoparticles and Hormesis. Dose-Response 8, 501-517 (2010).
- 52. Tang, J. et al. Graphene Oxide—Silver Nanocomposite as a Highly Effective Antibacterial Agent with Species-Specific Mechanisms. ACS Appl. Mater. Interfaces 5, 3867—3874 (2013).
- 53. Kirk, T. K., Schultz, E., Connors, W. J., Lorenz, L. F. & Zeikus, J. G. Influence of Culture Parameters on Lignin Metabolism by *Phanerochaete chrysosporium. Arch. Microbiol.* 117, 277–285 (1978).
- 54. Luo, Y. H. et al. Cadmium-based Quantum Dot Induced Autophagy Formation for Cell Survival via Oxidative Stress. Chem. Res. Toxicol. 26, 662–673 (2013).
- 55. Chen, A. et al. Plasma Membrane Behavior, Oxidative Damage, and Defense Mechanism in *Phanerochaete chrysosporium* under Cadmium Stress. *Process Biochem.* 49, 589–598 (2014).

Acknowledgements

This study was financially supported by the National Natural Science Foundation of China (51579099, 51178171, 51521006 and 51508186), the Program for Changjiang Scholars and Innovative Research Team in University (IRT-13R17).

Author Contributions

The authors Z.G., G.C. and L.L. implemented the assays and wrote the paper; the author G.Z. participated in the paper writing; the data were mainly processed by the author Z.H. and the authors A.C. and L.H. polished this paper.

Additional Information

Supplementary information accompanies this paper at http://www.nature.com/srep

Competing financial interests: The authors declare no competing financial interests.

How to cite this article: Guo, Z. *et al.* Activity Variation of *Phanerochaete chrysosporium* under Nanosilver Exposure by Controlling of Different Sulfide Sources. *Sci. Rep.* **6**, 20813; doi: 10.1038/srep20813 (2016).

This work is licensed under a Creative Commons Attribution 4.0 International License. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in the credit line; if the material is not included under the Creative Commons license, users will need to obtain permission from the license holder to reproduce the material. To view a copy of this license, visit http://creativecommons.org/licenses/by/4.0/